BRITISH STANDARD

BS EN ISO 6468:1997
BS 6068-2.57: 1997

Water quality — Determination of certain organochlorine insecticides, polychlorinated biphenyls and chlorobenzenes — Gas chromatographic method after liquid-liquid extraction

The European Standard EN ISO 6468:1996 has the status of a British Standard

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ICS 13.060.01

Committees responsible for this British Standard

The preparation of this British Standard was entrusted by Technical Committee EH/3, Water quality, to Subcommittee EH/3/2, Physical, chemical and biochemical methods, upon which the following bodies were represented:

British Agrochemicals Association Ltd.
British Ceramic Research
British Gas plc
British Soft Drinks Association Ltd.
Chartered Institution of Water and Environmental Management
Chemical Industries' Association
Convention of Scottish Local Authorities
Environment Agency
GAMBICA (BEAMA) Ltd.
Industrial Water Society
Laboratory of the Government Chemist
Royal Society of Chemistry
Soap and Detergent Industry Association
Society of Chemical Industry
Swimming Pool and Allied Trades Association Ltd.
Water Companies Association
Water Reseach Centre
Water Services Association of England and Wales

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Contents

National foreword

This British Standard has been prepared by Subcommittee EH/3/2 and is the English language version of EN ISO 6468:1996 *Water quality — Determination of certain organochlorine insecticides, polychlorinated biphenyls and chlorobenzenes — Gas chromatographic method after liquid-liquid extraction* published by the European Committee for Standardization (CEN). It is identical with ISO 6468:1996, published by the International Organization for Standardization (ISO).

BS EN ISO 6468 is one of a series of standards on water quality, others of which have been, or will be, published as Sections of BS 6068. This standard has therefore been given the secondary identifier BS 6068-2.57. BS 6068 comprises the following Parts, which are divided into Sections.

— *Part 0: Introduction;*

— *Part 1: Glossary;*

— *Part 2: Physical, chemical and biochemical methods;*

— *Part 3: Radiological methods;*

— *Part 4: Microbiological methods;*

— *Part 5: Biological methods;*

— *Part 6: Sampling;*

— *Part 7: Precision and accuracy.*

NOTE The tests described in this British Standard should only be carried out by suitably qualified persons with an appropriate level of chemical expertise. Standard chemical procedures should be followed throughout.

Cross-references

Publication referred to	Corresponding British Standard
EN 25667-1:1993 (ISO 5667-1:1980)	BS 6068 *Water quality* Part 6 *Sampling* Section 6.1:1981 *Guidance on the design of sampling programmes (also numbered BS EN 25667-1:1994)*
EN 25667-2:1993 (ISO 5667-2:1991)	Section 6.2:1991 *Guidance on sampling techniques (also numbered BS EN 25667-2:1993)*

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Summary of pages

This document comprises a front cover, an inside front cover, pages i and ii, the EN title page, page 2, the ISO title page, page ii, pages 1 to 24, an inside back cover and a back cover.

This standard has been updated (see copyright date) and may have had amendments incorporated. This will be indicated in the amendment table on the inside front cover.

EUROPEAN STANDARD

NORME EUROPÉENNE

EUROPÄISCHE NORM

EN ISO 6468

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English version

Water quality — Determination of certain organochlorine insecticides, polychlorinated biphenyls and chlorobenzenes — Gas chromatographic method after liquid-liquid extraction (ISO 6468:1996)

Qualité de l'eau — Dosage de certains insecticides organochlorés des polychlorobiphényles et des chlorobenzènes — Méthode par chromatographie en phase gazeuse après extraction liquide-liquide (ISO 6468:1996)

Wasserbeschaffenheit — Bestimmung bestimmter Organochlorinsektizide, Polychlorbiphenyle und Chlorbenzole — Gaschromatographisches Verfahren nach Flüssig-Flüssig-Extraktion (ISO 6468:1996)

This European Standard was approved by CEN on 1996-11-08. CEN members are bound to comply with the CEN/CENELEC Internal Regulations which stipulate the conditions for giving this European Standard the status of a national standard without any alteration.

Up-to-date lists and bibliographical references concerning such national standards may be obtained on application to the Central Secretariat or to any CEN member.

The European Standards exist in three official versions (English, French, German). A version in any other language made by translation under the responsibility of a CEN member into its own language and notified to the Central Secretariat has the same status as the official versions.

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CEN

European Committee for Standardization
Comité Européen de Normalisation
Europäisches Komitee für Normung

Central Secretariat: rue de Stassart 36, B-1050 Brussels

Foreword

The text of the International Standard ISO 6468:1996 has been prepared by Technical Committee ISO/TC 147 "Water quality" in collaboration with Technical Committee CEN/TC 230 "Water analysis", the secretariat of which is held by DIN.

This European Standard shall be given the status of a national standard, either by publication of an identical text or by endorsement, at the latest by June 1997, and conflicting national standards shall be withdrawn at the latest by June 1997.

According to the CEN/CENELEC Internal Regulations, the national standards organizations of the following countries are bound to implement this European Standard: Austria, Belgium, Denmark, Finland, France, Germany, Greece, Iceland, Ireland, Italy, Luxembourg, Netherlands, Norway, Portugal, Spain, Sweden, Switzerland and the United Kingdom.

Endorsement notice

The text of the International Standard ISO 6468:1996 was approved by CEN as a European Standard without any modification.

NOTE Normative references to International Standards are listed in Annex ZA (normative).

Contents

INTERNATIONAL STANDARD

ISO 6468

First edition
1996-12-15

Water quality — Determination of certain organochlorine insecticides, polychlorinated biphenyls and chlorobenzenes — Gas chromatographic method after liquid-liquid extraction

Qualité de l'eau — Dosage de certains insecticides organochlorés, des polychlorobiphényles et des chlorobenzènes — Méthode par chromatographie en phase gazeuse après extraction liquide-liquide

Reference number
ISO 6468:1996(E)

Foreword

ISO (the International Organization for Standardization) is a worldwide federation of national standards bodies (ISO member bodies). The work of preparing International Standards is normally carried out through ISO technical committees. Each member body interested in a subject for which a technical committee has been established has the right to be represented on that committee. International organizations, governmental and non-governmental, in liaison with ISO, also take part in the work. ISO collaborates closely with the International Electrotechnical Commission (IEC) on all matters of electrotechnical standardization.

Draft International Standards adopted by the technical committees are circulated to the member bodies for voting. Publication as an International Standard requires approval by at least 75 % of the member bodies casting a vote.

International Standard ISO 6468 was prepared by Technical Committee ISO/TC 147, *Water quality,* Subcommittee SC 2, *Physical, chemical, biochemical methods.*

Annex A forms an integral part of this International Standard. Annex B to Annex A are for information only.

WARNING AND SAFETY PRECAUTIONS — This method makes use of flammable and toxic organic solvents. Observe the safety regulations in effect.

The electron-capture detector (ECD) contains radionuclides. Adequate safety precautions and legal requirements must be observed.

The halogenated hydrocarbons and chloropesticides, used for the preparation of the calibration standards are toxic. Therefore, the safety regulations pertaining must be strictly observed.

1 Scope

This International Standard describes a method for determining certain organochlorine insecticides, polychlorinated biphenyls (PCBs) and chlorobenzenes (except the mono- and dichlorobenzenes) in drinking waters, ground waters, surface waters and waste waters.

The method is applicable to samples containing up to 0,05 g/l of suspended solids. In the presence of organic matter, suspended matter and colloids, interferences are more numerous and consequently the detection limits are higher.

The method described in this International Standard only gives information on specific PCB compounds but no information on the level of total PCBs.

According to the types of compounds to be detected and the source of the water, the detection limits given in Table 1 are applicable for the method described in this International Standard, with waters of low organic contents.

Given the very low concentrations normally present in the waters, the problem of contamination is extremely important. The lower the level measured, the more precautions have to be observed; below concentrations of 10 ng/l, special care is necessary.

2 Normative references

The following standards contain provisions which, through reference in this text, constitute provisions of this International Standard. At the time of publication, the editions indicated were valid. All standards are subject to revision, and parties to agreements based on the International Standard are encouraged to investigate the possibility of applying the most recent editions of the standards indicated below. Members of IEC and ISO maintain registers of currently valid International Standards.

ISO 5667-1:1980, *Water quality — Sampling — Part 1: Guidance on the design on sampling programmes.*

ISO 5667-2:1991, *Water quality — Sampling — Part 2: Guidance on sampling techniques.*

3 Principle

Liquid-liquid extraction of organochlorine insecticides, chlorobenzenes and PCBs by an extraction solvent. After the concentration of the components with low volatility and after any clean-up steps which may be necessary, the sample extracts are analysed by gas chromatography, using an electron-capture detector.

Any substance capable of producing a response on the electron-capture detector, at a retention time indistinguishable from any compound of interest, will interfere. In practice, many potentially interfering substances will be removed during the extraction and clean-up procedures.

NOTE 1 In general, the use of two capillary columns of different polarity is sufficient for the organochlorine compounds analysed according to this International Standard. The results so calculated should be considered as the maximum concentrations, possibly still influenced by coeluting substances. It is possible that there will be cases where a more definite identification is required.

4 Reagents and materials

All reagents shall be sufficiently pure to not give rise to significant interfering peaks in the gas chromatograms of the blanks. The purity of reagents used in the procedure shall be checked by blank determinations (**7.6**).

NOTE 2 Commercial "pesticide grade" solvents are available. The use of these products is recommended only after verifying their quality. The quality of a solvent is checked by evaporation of about 200 ml down to 1 ml and analysis of the concentrate to determine the compounds subsequently analysed. The solvent should be considered acceptable if it does not give any detectable interfering peaks in the chromatogram for the substance of interest.

4.1 *Water purified,* for example, using ion-exchange or carbon-column adsorption.

4.2 *Extraction solvent*

Hexane, petroleum ether or heptane are suitable.

NOTE 3 Any other solvents meeting the requirements of **8.3** (recovery rate ⩾ 60 %) may be used.

4.3 *Sodium sulfate* (Na_2SO_4), anhydrous.

Heat a portion of about 250 ml to 300 ml of sodium sulfate powder at 500 °C ± 20 °C for 4 h ± 30 min, cool to about 200 °C in a muffle furnace and then to ambient temperature in a desiccator containing magnesium perchlorate or an equivalent alternative.

4.4 *Decane* ($C_{10}H_{22}$) or *dodecane* ($C_{12}H_{26}$), or any keeper which is not detected by the electron-capture detector.

Table 1 — Detection limits

Acronyms	Chemical names (IUPAC)	Detection limits
Organochlorine insecticides:		
HCH **Lindane**	1, 2, 3, 4, 5, 6-hexachlorocyclohexane, five stereoisomers: alpha-HCH, beta-HCH, gamma-HCH, delta-HCH, epsilon-HCH	
o,*p*′-DDE	1, 1-dichloro-2-(2-chlorophenyl 1)-2-(4-chlorophenyl)ethylene	
p,*p*′-DDE	1, 1-dichloro-2, 2-bis(4-chlorophenyl)ethylene	1 ng/l
o,*p*′-TDE	1, 1-dichloro-2-(2-chlorophenyl)-2-(4-chlorophenyl)ethane	to
p,*p*′-TDE	(= o,*p*′-DDD)	10 ng/l
o,*p*′-DDT	1, 1-dichloro-2, 2-bis(4-chlorophenyl)ethane (= *p*,*p*′-DDD)	depending
p,*p*′-DDT	1, 1, 1-trichloro-2-(2-chlorophenyl)-2-(4-chlorophenyl)ethane	on the
Methoxychlor	1, 1, 1-trichloro-2, 2-bis(4-chlorophenyl)ethane	compound
Aldrin	1, 1, 1-trichloro-2, 2-bis(4-methoxyphenyl)ethane	
Dieldrin	(1*R*, 4*S*, 4a*S*, 5*S*, 8*R*, 8a*R*)-1, 2, 3, 4, 10, 10-hexachloro-1, 4, 4a, 5, 8, 8a-hexahydro-1, 4 : 5, 8-dimethanonaphthalene	
Endrin	(1*R*, 4*S*, 4a*S*, 5*R*, 6*R*, 7*S*, 8*S*, 8a*R*)-1, 2, 3, 4, 10, 10-hexachloro-1, 4, 4a, 5, 6, 7, 8, 8a-octahydro-6, 7-epoxy-1, 4 : 5, 8-dimethanonaphthalene	
Heptachlor[a]	(1*R*, 4*S*, 4a*S*, 5*S*, 6*S*, 7*R*, 8*R*, 8a*R*)-1, 2, 3, 4, 10, 10-hexachloro-1, 4, 4a, 5, 6, 7, 8, 8a-octahydro-6, 7-epoxy-1, 4 : 5, 8-dimethanonaphthalene	
Heptachlor-epoxide	1, 4, 5, 6, 7, 8, 8-heptachloro-3a, 4, 7, 7a-tetrahydro-4, 7-methanoindene[a]	
Endosulfan[ab]	1, 4, 5, 6, 7, 8, 8-heptachloro-2, 3-epoxy-3a, 4, 7, 7a-tetrahydro-4, 7-methanoindane	
	1, 4, 5, 6, 7, 7, 7-hexachloro-8, 9, 10-trinorborn-5-en-2, 3-ylene-dimethyl-enesulfite: alpha-Endosulfan, beta-Endosulfan	
Chlorobenzenes:		
TrCB	trichlorobenzene	1 ng/l
TeCB	tetrachlorobenzene	to 10 ng/l
PeCB	pentachlorobenzene	depending
HCB	hexachlorobenzene	on the
PCNB (Quintozene)	pentachloronitrobenzene	compound
Polychlorinated biphenyls:		
PCB 28	2,4,4′-trichlorobiphenyl	1 ng/l
PCB 52	2,2′,5,5′-tetrachlorobiphenyl	to
PCB 101	2,2′,4,5,5′-pentachlorobiphenyl	50 ng/l
PCB 138	2,2′,3,4,4′,5′-hexachlorobiphenyl	depending
PCB 153	2,2′,4,4′,5,5′-hexachlorobiphenyl	on the
PCB 180	2,2′,3,4,4′,5,5′-heptachlorobiphenyl	compound
PCB 194	2,2′,3,3′,4,4′,5,5′-octachlorobiphenyl	

[a] The analysis of α and β — endosulfan as well as heptachlor requires special care due to its low stability.

[b] The name "endosulfan" is not acceptable for use in Italy, as it is in conflict with a trade mark registered there.

4.5 *Dry alumina*

Heat a batch of inert alumina, containing particles of size 50 µm to 200 µm and of maximum mass 500 g, at 500 °C ± 20 °C for 4 h ± 30 min on a silica dish in a muffle furnace. Cool to about 200 °C in the furnace and then to ambient temperature in a desiccator. Store in a sealed glass container.

4.6 *Deactivated alumina*

Weigh a portion of dry alumina (**4.5**) into a sealable all-glass container and add 7 % ± 0,2 % (*m*/*m*) of water (**4.1**). Seal and agitate for at least 2 h to ensure uniformity. Store in a sealed glass container.

Once the seal has been broken, storage time is normally about one week. After the maximum storage time, reprocess batches as described in **4.5** and this subclause.

4.7 *Alumina/silver nitrate*

Dissolve 0,75 g ± 0,01 g of silver nitrate in 0,75 ml ± 0,01 ml of water (**4.1**) using a microburette. Add 4,0 ml ± 0,2 ml of acetone followed by 10 g ± 0,2 g of deactivated alumina (**4.6**). Mix thoroughly by shaking in an open-topped conical flask, protected from light. Allow the acetone to evaporate at room temperature and prevent condensation, for example by warming with the hand.

Store in the dark and use within 4 h after preparation.

4.8 *Silica gel,* of particle size 63 µm to 200 µm, heated at 500 °C ± 30 °C in batches not larger than 500 g, for about 14 h. Cool to about 200 °C in the furnace and then to ambient temperature in a sealed flask which is placed in a desiccator without desiccant. Use this material within one week. Deactivate the silica gel by weighing a suitable quantity of silica and adding 3 % (*m*/*m*) of water (**4.1**). Agitate for at least 2 h to ensure uniformity and store in a sealed glass container.

The deactivated silica gel shall be used within 24 h.

4.9 *Toluene*

4.10 *Diethylether,* free from peroxides.

4.11 *Anti-bumping granules,* washed with acetone.

4.12 *Standard stock solutions*

Pure or certified standards of organochlorine insecticides, chlorobenzenes, and PCBs shall be used for the preparation of standard stock solutions.

NOTE 4 Suitable solvents for the preparation of standard stock solutions are acetone, pentane, hexane, dimethylbenzene or isooctane.

The containers containing the solutions shall be marked or weighed so that any evaporation losses of the solvent may be recognized. The solutions shall be stored in volumetric flasks with ground-glass stoppers at a temperature of 4 °C in the dark. Prior to use, they shall be brought to ambient temperature and the level of solvent shall be adjusted, if necessary.

NOTE 5 A convenient concentration of standard stock solution is obtained by weighing 50 mg of each determinand and dissolving it in 100 ml of the solvent.

The solution is stable for about 1 year.

4.13 *Intermediate standard solutions*

Prepare intermediate standard solutions by a suitable dilution of the stock solution (**4.12**) with the extraction solvent (**4.2**).

A typical value is 10 µg/ml.

Store the intermediate standard solutions at about 4 °C in the dark. These solutions are stable for six months.

4.14 *Working standard solutions*

Prepare at least five different concentrations by suitable dilutions of the intermediate standard solutions (**4.13**) with the extraction solvent (**4.2**).

Suitable concentrations are in the nanograms per millilitre range.

Store the solutions at about 4 °C in the dark. These solutions are stable for at least one month.

4.15 *Cotton wool or glass wool,* washed with extraction solvent.

4.16 *Water-miscible solvent*

NOTE 6 Acetone, methanol or dimethylformamide may be used.

5 Apparatus

5.1 *Gas chromatograph,* with an electron-capture detector (ECD) and suitable for use with capillary columns. This shall be operated in accordance with the manufacturer's instructions. On-column or glass-lined injection systems can be used. The oven shall be suitable for isothermal and temperature-programmable operation.

5.2 *Capillary columns,* glass or fused-silica capillaries, with an inside diameter of less than 0,4 mm and a length of 25 m to 60 m, coated with stationary phases capable of separating the compounds of interest.

Annex B provides examples of gas chromatographic conditions (Table B.1, Table B.2 and Table B.3) and the corresponding gas chromatograms (Figure B.1 and Figure B.2).

5.3 *Separating funnels,* of nominal capacities 1 litre to 5 litres, with a glass tap washed by hexane or a polytetrafluoroethylene (PTFE) tap.

5.4 *High-speed stirrer and magnetic stirring bar,* washed with hexane and coated with polytetrafluoroethylene (PTFE).

5.5 *Microseparator,* see example in Figure C.1.

5.6 *Kuderna-Danish evaporator,* see example in Figure D.1.

5.7 *Snyder microcolumn*

5.8 *Rotary evaporator or any suitable system of evaporation*

5.9 *Column for drying the extract,* filled with 5 g to 7 g of sodium sulfate (**4.3**) giving a height of about 7 cm to 10 cm. For example, the dimensions are 10 mm internal diameter and 250 mm length (see Figure E.1).

5.10 *Column for the alumina-alumina/silver nitrate clean-up,* for example, the dimensions are 10 mm internal diameter and 250 mm length (see Figure E.1).

5.11 *Macrocolumn for the silica gel clean-up,* for example, the dimensions are 19 mm internal diameter and 400 mm length (see Figure E.1).

5.12 *Microcolumn for the silica gel clean-up,* for the dimensions see Figure F.1.

5.13 *Microlitre syringes*

5.14 *Miscellaneous glassware*

Laboratory glassware shall be cleaned using a cleaning agent (laboratory detergent) followed, for example, by either a treatment with chromium(VI)/sulfuric acid mixture, or peroxodisulfate/sulfuric acid mixture and subsequently washed by hexane or heated for at least 12 h at 200 °C, except for the calibrated glassware.

The efficiency of the treatment shall be experimentally checked at random by blank determinations to ensure that no interfering contamination has occurred.

6 Sampling and sample preparation

Take samples according to ISO 5667-1 and ISO 5667-2.

Collect the water samples in brown glass bottles cleaned as described in **5.14** (do not use plastics bottles) with ground-glass stoppers or with screw caps with PTFE liners, of nominal capacity 1 litre to 5 litres. Fill the bottles to 80 % to 90 %.

On sample collection, ensure that no interfering substances enter the water sample, and no losses of the determinands occur. This is especially important when using plastics tubing with the sampling apparatus. If necessary, it shall be proved by control tests that no losses by adsorption occur. Glass and stainless steel devices shall preferably be used.

Check the pH. If necessary, correct the pH immediately after collection in order to be in the range pH 5 to 7,5.

If endosulfan is to be determined, take a separate sample and keep it under acidic conditions (pH 2) until extraction.

Do not place samples in close proximity to the concentrated insecticide or PCB or chlorobenzene solutions. Store in the dark at a temperature of approximately 4 °C prior to extraction.

Ensure that all samples are extracted as soon as possible (preferably within 24 h) to avoid decomposition of the compounds after sampling.

Halogenated hydrocarbons of low volatility and organochlorine insecticides are relatively stable if transferred into an organic solvent. Therefore, it is permissible to store the dried solvent extracts in a refrigerator at 4 °C for up to two months. Evaporation of the solvent can still occur even under refrigeration. Extracts shall not be allowed to go to dryness and the volume of solvent shall be restored to the original amount before starting analysis.

7 Procedure

7.1 Sample pretreatment

Sample pretreatment is not normally necessary.

If the sample container is filled up to the ground-glass joint, shake and pour off 30 ml to 100 ml of the sample in order to obtain sufficient free volume for the subsequent addition of the solvent.

Measure the volume of the water to be extracted by weighing the bottle before extraction and after emptying.

7.2 Extraction and separation

Use either of these two procedures for extraction and separation:

— extraction in the sample container and separation in a separating funnel (**7.2.1**);

— extraction in the sample container with a magnetic stirrer or a high-speed stirrer and separation by a microseparator (**7.2.2**).

NOTE 7 Depending on the method used, varying recoveries and reproducibilities may be obtained. The yields of the selected method should be checked by the laboratory (**8.3**).

It is recommended to perform the extraction in the sample container. Usually, a sample volume of about 1 litre is used.

7.2.1 ***Extraction by shaking the sampling bottle and separation in a separating funnel***

Add 30 ml of the extraction solvent (**4.2**) to the sample (**7.1**) and shake for at least 10 min.

Transfer to a separating funnel of suitable capacity (**5.3**) and allow the phases to separate.

Run the lower aqueous phase back into the sample container. Repeat the extraction twice with 20 ml to 30 ml of the extraction solvent (**4.2**).

Dry the extract using one of the following procedures:

— Pass the extract through a drying column (**5.9**) containing anhydrous sodium sulfate (**4.3**), previously washed with the solvent (**4.2**) and collect the eluate in the evaporating vessel.

NOTE 8 It is advisable to wash the column with a further 10 ml to 20 ml portion of the solvent (**4.2**) to obtain a better recovery. Collect the washings in the evaporating vessel.

Or

— Add anhydrous sodium sulfate (**4.3**) to the flask. Shake for 1 min. Leave for 5 min and decant the extract into the concentration apparatus. The sodium sulfate is washed with a further 10 ml to 20 ml of solvent (**4.2**) and the washings added to the evaporating vessel.

Or

— Freeze the extract at – 18 °C for 2 h. The solvent extract is decanted from the ice and transferred to the evaporating vessel. The ice is washed with a further 10 ml of solvent (**4.2**) and the washings are added to the evaporating vessel.

7.2.2 ***Extraction with a magnetic or a high-speed stirrer and separation in a microseparator***

Add 20 ml to 30 ml of the extraction solvent (**4.2**) to the sample (**7.1**).

With a magnetic stirrer and a stirring bar (**5.4**), stir for at least 10 min, at a speed of at least 1 000 r/min (the solvent needs to be dispersed finely in the water) keeping the sample covered, and then allow the phases to separate. Alternatively, if a high-speed stirrer (**5.4**) is used, stir for 2 min while keeping the sample covered at a temperature of 4 °C and allow the phases to separate.

Assemble the microseparator (**5.5**); pour purified water (**4.1**) into the funnel until the surface of the organic phase rises sufficiently for the extract to be withdrawn with a pipette.

Dry the extract as described in **7.2.1**.

7.3 Concentration of the extract

Concentrate the combined dried extracts from either **7.2.1** or **7.2.2** by either of the procedures described in **7.3.1** or **7.3.2** or by any other suitable system (**5.8**). Ensure that no significant losses of the more volatile determinands of interest occur.

7.3.1 ***Concentration using a Kuderna-Danish evaporator***

Good detection limits can be obtained by evaporating the sample extract to a small volume with the Kuderna-Danish evaporator (**5.6**) and a Snyder microcolumn (**5.7**) as follows.

Collect the dried extract in a Kuderna-Danish evaporator.

Add two anti-bumping granules (**4.11**) and evaporate to 5 ml ± 1 ml on a steam bath. Further concentrate the extract to less than 1 ml using a Snyder microcolumn or a gentle stream of clean inert gas (e.g. nitrogen) with a tube placed in a warm water bath (not exceeding 40 °C).

NOTE 9 No further precautions are necessary if the extract is evaporated with this apparatus to a final volume of not less than 0,5 ml. If a smaller final volume is required, it is recommended to use a keeper (**4.4**) in order to avoid significant losses. Decane or dodecane may be used as keepers because they are not detected by the electron-capture detector. 0,1 ml of a solution containing 20 g/l of decane or dodecane in hexane are added to the extract to be concentrated.

7.3.2 ***Concentration using a rotary evaporator***

Concentrate the extract in a tapered flask, or preferably, in a tapered flask with an ampoule extension on a rotary evaporator (**5.8**) to not less than 0,6 ml at a constant vacuum of greater than 340 mbar. A Kuderna-Danish evaporation flask (**5.6**) is mounted between the evaporating vessel and the rotary evaporator.

Place the evaporating vessel with the solvent extract in an unheated water bath or, for higher boiling extractants, in a water bath at a temperature not exceeding 50 °C. When the concentration is finished, quantitatively transfer the extract into a 1 ml measuring flask. Carefully rinse the walls of the evaporating vessel with a small volume of solvent (**4.2**). Transfer the rinsings to the measuring flask and fill up to volume with the solvent.

7.4 Gas chromatography

For extracts of samples from clean waters, perform gas chromatographic analysis at this stage without further clean-up.

If the analysis has to be performed with a purification step, proceed to **7.5**.

Set up the gas chromatograph (**5.1**), fitted with an electron-capture detector and equipped with a suitable column (**5.2**) according to the instructions of the manufacturer, and ensure it is in a stable condition.

Inject the extract (usually between 1 µl and 10 µl but the same volume as that used for calibration) into the gas chromatograph and run a chromatogram.

Compare the gas chromatogram obtained to those of the standard solutions (see clause **8**).

Evaluate the gas chromatogram qualitatively and quantitatively (see clause **9**).

The requirements applicable to the extent of the measurements, and the calibration, evaluation and calculation techniques to be used, are described in clause **8**. The gas chromatogram obtained is checked for overlapping occurring at the locations of the retention times of the determinands of interest. If interfering peaks are present, one of the purification methods described in **7.5** shall be applied. Otherwise, identify and quantify according to clause **9**.

7.5 Clean-up and separation

Applying the procedure described in **7.2** may lead to coextraction of relatively polar and/or other undesired substances, which are likely to interfere by the appearance of unknown peaks overlapping the pesticide peaks.

NOTE 10 Treatment by column chromatography may help to eliminate some of the substances. However, this method cannot be considered as an absolute system.

Use one or both of the following procedures:

— clean-up on an alumina-alumina/silver nitrate column, for purification to remove polar compounds (**7.5.1**);

— clean-up on a silica gel column, for separation of PCB from most insecticides (**7.5.2**).

NOTE 11 The quality of each batch of columns should be checked with standard solutions.

7.5.1 *Clean-up on alumina-alumina/silver nitrate column*

Carry out the purification on an alumina-alumina/silver nitrate column as described in **7.5.1.1** and **7.5.1.2**. If interference persists, the additional procedure described in Annex A may be carried out.

NOTE 12 Some compounds, for example endosulfan, may be retained on the column.

7.5.1.1 *Preparation of the column*

Place 15 ml ± 1 ml of the extraction solvent (**4.2**) in the column (**5.10**), then add 1,0 g ± 0,2 g of alumina/silver nitrate (**4.7**) and allow to settle while tapping gently. Then add 2,0 g ± 0,2 g of alumina (**4.6**) and again allow to settle while tapping gently. Add a sufficient amount of sodium sulfate (**4.3**) to produce a 5 mm layer on top of the column. Prepare the column immediately before use.

7.5.1.2 *Purification*

Prepare an alumina-alumina/silver nitrate column as described in **7.5.1.1**. Run off the surplus of the extraction solvent (**4.2**). When the solvent level reaches the top of the column, add the concentrated sample extract (see **7.3**). Wash the sample vessel with 2 ml ± 0,5 ml of extraction solvent and add the washings to the column. Elute the column with 30 ml ± 1 ml of extraction solvent. Collect and concentrate the extract as described in **7.3** and then perform the gas chromatographic analysis according to **7.4**.

During addition to the column, do not allow the meniscus of the solvent (**4.2**) to fall below the surface of the alumina. If the alumina/silver nitrate column blackens along its entire length, prepare a fresh column (see **7.5.1.1**) and repeat the purification. If total blackening is a common occurrence, larger columns may be used but additional solvent will be required for elution.

7.5.2 *Clean-up on silica gel*

7.5.2.1 *Preparation of the column*

Choose a chromatography column (**5.12**) as shown in Figure F.1 in Annex F. [Initially without the solvent reservoir (Figure E.2) attached.] Plug the column temporarily with a rubber cap at the lower end, and fill it with extraction solvent (**4.2**).

Insert a plug of glass wool (**4.15**) close to the lower end.

Suspend 1 g of silica gel (**4.8**) in the extraction solvent (**4.2**) in a small beaker.

Transfer the suspension to the chromatography column with the aid of a pipette.

Let the silica gel settle down during constant vibration of the column, to produce a dense layer. Otherwise, the sodium sulfate which is placed onto the silica gel will move into the silica gel layer.

Remove the rubber cap.

Carry out the following steps, including the steps described in **7.5.2.2**, without interuption as soon as the column starts dripping continuously.

Place 0,2 g of sodium sulfate (**4.3**) onto the layer of silica gel. Attach the solvent reservoir to the column and rinse the system with 5 ml of solvent (**4.2**).

Once again, remove the solvent reservoir as soon as the level of solvent has moved down to the column section of the apparatus and follow the steps described in **7.5.2.2** immediately.

NOTE 13 Alternatively, dry packed and/or commercially available disposable columns may be used, if they are found to be equally suitable.

7.5.2.2 *Clean-up and separation*

Add 100 µl of the sample extract onto the column with the aid of a 100 µl syringe, just before the meniscus of the solvent has reached the sodium sulfate layer.

NOTE 14 The flow rate should be about 1 to 2 drops per second.

NOTE 15 Depending on the concentration of organochlorine compounds in the sample, it is recommended that at least 1/10 of the whole sample extract be taken for the clean-up. This means that the sample extract has to be concentrated to a volume of 1 ml or less by the methods described in **7.3**, prior to the clean-up.

Attach the solvent reservoir again (see **7.5.2.1**) and add 5 ml of extraction solvent (**4.2**).

For the acceleration of the chromatography process, connect a pressurized inert gas supply (e.g. nitrogen) at a pressure of about 25 mbar.

Collect the first fraction in a graduated Kuderna-Danish vessel. When the meniscus of the solvent has reached the sodium sulfate layer, add additional solvent. After disconnecting from the pressurized gas supply, repeat the steps in the following order:

— second fraction: 2,5 ml of solvent (**4.2**);

— third fraction: 2,5 ml of solvent;

— fourth fraction: 8 ml of solvent;

— fifth fraction: 8 ml of solvent/toluene (**4.9**) (95 : 5) (*V*/*V*);

— sixth fraction: 16 ml of solvent/toluene (90 : 10) (*V*/*V*);

— seventh fraction: 8 ml of solvent/diethylether (**4.10**) (99,5 : 0.5) (*V*/*V*).

Before concentrating, combine the fractions as appropriate.

Concentrate the fractions collected as described in **7.3** and then perform a gas chromatographic analysis according to **7.4**.

NOTE 16 Table G.1 in Annex G gives a typical example covering the elution sequence of 27 compounds and of their recoveries with the macrocolumn for the silica gel clean-up (**5.11**), including a subsequent concentration with the rotary evaporator procedure.

7.6 Blank determination

Carry out the complete procedure (pretreatment, extraction, concentration, clean-up, gas chromatographic analysis) using a sample of pure water (**4.1**).

If the blank value is unreasonably high, i.e. greater than 10 % of the lowest value for any of the compounds of interest, carry out a step-by-step examination of the procedure and eliminate the cause.

8 Calibration

Initially, it is necessary to determine the recovery using the following methods.

a) Calibration by direct injection of solvent standard solutions (**8.1**).

This gives information on the linear working range of the detector, retention times and relative responses of the determinands.

b) Calibration of the overall procedure (**8.2**) using water samples (preferably of the same type as those being analysed), which are spiked and extracted and, if necessary, cleaned-up.

The data obtained from a) are compared with those from b) in order to calculate the recovery (**8.3**) of each determinand.

Carry out the daily recalibration (**8.4**) with solvent standard solutions according to a) or with spiked water extracts according to b).

Table 2 gives an explanation of the subscripts used in the equations and in the explanations of symbols after the equations.

Table 2 — Explanation of the subscripts used in the symbols

Index	Meaning
i	Identity of the determinand
e	Measured value in calibration
g	Entire procedure

8.1 Calibration by external standard, not using the overall procedure

Inject volumes in the range of 1 µl to 10 µl of the working standard solutions (**4.14**) into the gas chromatograph.

Measure the gas chromatographic signals for each substance (peak heights or peak areas or area integration units, respectively) and calculate the concentrations.

For a graphic presentation of the calibration curve, plot the respective measured values, y_{ie}, on the ordinate against the respective mass concentrations, ρ_{ie}, of the substance i (e.g. in the solvent) on the abscissa.

The injection volume used for calibration and for the measurement of the sample solutions shall be kept constant.

The series of measured values thus obtained shall be used to establish the linear regression function as follows:

$$y_{ie} = m_i \cdot \rho_{ie} + b_i \quad ...(1)$$

where

y_{ie} is the dependent variable: measured response of the substance i, dependent on ρ_{ie} (its unit depends on the evaluation, e.g. area value);

ρ_{ie} is the independent variable: mass concentration of the substance i (external standard) in the calibration solution, in nanograms per microlitre;

m_i is the slope of the calibration curve of the substance i (its unit depends on the evaluation, e.g. area value × µl/ng);

b_i is the intercept of the calibration curve on the ordinate (its unit depends on the evaluation, e.g. area value). As a rule, the intercept is very small. If large intercepts occur, omit the highest concentration(s) of the standard(s) and recalculate the linear regression function. This should reduce the value of the intercept. If not, the gas chromatographic system and the evaluation system should be checked.

8.2 Calibration of the overall procedure using an external standard

For each compound, a separate calibration graph (via the overall procedure) shall be established, consisting of at least five points. It is permissible to examine several compounds in one calibration experiment.

To calibrate the entire procedure, prepare aqueous solutions by spiking water (**4.1**) with the compounds to be determined in an individual concentration range within the linear dynamic range of the detector, as follows.

8.2.1 *Preparation of the spiked aqueous standard solutions*

To a 100 ml graduated flask, containing about 90 ml of water-miscible solvent (**4.16**) using a microlitre syringe (**5.13**), add defined quantities of the standard stock solutions (**4.12**) of each determinand, under the surface of the solvent.

Immediately dilute to volume with the water-miscible solvent (**4.16**).

Stopper the flask with its ground-glass stopper and cautiously shake the solution.

Calculate the respective concentration of each substance added.

The stock solution prepared in this way can be stored at a temperature of about 4 °C in the dark for several weeks. Prior to use, equilibrate at room temperature for at least 15 min.

Prepare at least five spiked aqueous standard solutions covering (depending on the compounds) the range 1 ng/l to 200 ng/l, by adding different volumes of this stock solution to water (**4.1**).

For blank measurements, to one bottle of water (**4.1**), add the same quantity of solvent as that used for the preparation of the spiked aqueous standard solutions.

Use the quantities such that the volume added is as small as possible (< 1 ml/l of water), in order to minimize any effect on the partition equilibrium.

Prepare the spiked aqueous standard solutions on the day of use.

8.2.2 *Calibration curve*

Extract and concentrate these spiked aqueous standard solutions as described in **7.2** and **7.3**.

Inject the extract of the blank into the gas chromatograph, and then the calibration solutions with concentrations ρ_{ieg} in ascending order. Measure the peak values y_{ieg} of the calibration samples.

Calculate a regression function for each substance using the pairs of values y_{ieg} and ρ_{ieg}:

$$y_{ieg} = m_{ig} \cdot \rho_{ieg} + b_{ig} \quad ...(2)$$

where

y_{ieg} is the dependent variable: measured response of the substance i during calibration, dependent on ρ_{ieg}, (its unit depends on the evaluation, e.g. area value);

ρ_{ieg} is the independent variable: mass concentration of the substance i in the calibration solution, expressed in micrograms per litre;

m_{ig} is the slope of the calibration curve of the substance i, often referred to as f_i (its unit depends on the evaluation, e.g. area value × l/µg);

b_{ig} is the intercept of the calibration curve on the ordinate (its unit depends on the evaluation, e.g. area value).

Plot the reference functions in a diagram with the ordinate as the specific measured signals of the substance i, y_{ieg}, and the abscissa as the mass concentrations, ρ_{ieg}, of the substance i in the spiked aqueous calibration solution. With the aid of the calibration curve, define the working range of the procedure.

8.3 Determination of the recovery

By means of the calibration procedure according to **8.1** and **8.2**, determine the specific mean recovery A_i, for the substance i using equation (3):

$$A_i = \frac{m_{ig} \cdot F_V}{m_i} = \frac{m_{ig} \cdot V_E \cdot f}{m_i \cdot V_p} \quad ...(3)$$

A_i is the mean recovery for the substance i (dimensionless);

m_i is as defined in **8.1**;

m_{ig} is as defined in **8.2.2**;

f is a conversion factor (here f = 1 000);

F_V is the ratio of the volume of extraction liquid and sample. This factor has to be calculated while taking into account sample volume, extractant volume, dilution factors (if applicable) and injection volumes if they differ from those used for calibration. The following equation applies:

$$F_V = \frac{V_E \cdot f}{V_P} \quad ...(4)$$

where

V_E is the extractant volume, in millilitres;

V_P is the sample volume, in millilitres.

The recovery thus obtained is valid only for the experimental conditions used.

NOTE 17 Equation (3) is valid if b_i and b_{ig} are relatively small and if calibration according to equations (1) and (2) refers to the same range of concentration (in the extract and in the standard relation), for example comparable values for y_{ie} and y_{ieg}.

A high recovery is an essential prerequisite for a good precision and accuracy of the analytical result. Variations of these values will indicate problems in extraction and preparation of standards. The recovery depends on determinands and is generally greater than 60 %. If not, the procedure should be checked.

Recovery values obtained from different laboratories are given in Table H.1.

8.4 Recalibration

For routine recalibration of the method, it is essential to work within the previously established linear range (**8.1** or **8.2**). This shall be updated regularly, especially when contaminated samples such as sewage or trade effluents are analysed, as these may affect the detector and hence the linear range.

The minimum requirement for daily recalibration shall be injections of two solvent standard solutions (**4.14**) or two spiked water extracts (**8.2**). The concentration of the first solution shall be about 20 % of the selected linear working range and the concentration of the second solution about 80 % of this range.

Calculate a regression function.

Compare this function to the previously established calibration curve (**8.1** or **8.2**). If the values are within the range of the confidence limits of the previously established calibration curve (**8.1** or **8.2**), use the new calibration line for evaluation. If not, check the system and establish a complete new calibration curve.

9 Identification and evaluation

9.1 Identification of individual compounds

If, in the chromatogram of the sample extract run on a particular capillary column, no peak appears at the specific retention time of a substance, consider the compound as not being detected.

If a peak appears at a particular specific retention time of a substance, the presence of the compound requested is possible. The identity of this compound has to be confirmed.

Repeat the complete comparison procedure, using a capillary column (**5.2**) belonging to a different polarity group.

Normally, the reliability of the identification increases with increasing difference in the polarities of the columns applied. If the comparative study with two capillary columns of differing polarities reveals the presence of peaks at the expected specific retention times of a substance, consider the identity of the substance as highly probable.

NOTE 18 If necessary, mass spectrometry and chemical tests can be used for further confirmation.

9.2 Evaluation

9.2.1 *Evaluation using the (re)calibration according to 8.1*

Calculate the mass concentration, ρ_i, of the substance i in the water sample using equation (5), after solving equation (1) for the mass concentration, ρ_i.

$$\rho_i = \frac{y_i - b_i}{m_i} \qquad ...(5)$$

where

ρ_i is the mass concentration of the substance i in the water sample (uncorrected by recovery), expressed in micrograms per litre;

y_i is the measured value of the substance i in the extract of the water sample (on condition that the same procedure is applied as with the calibration and the sample measurement), (its unit depends on the evaluation, e.g. area value);

m_i is the slope of the calibration curve (**8.1** or **8.4**) of the substance i (its unit depends on the evaluation, e.g. area value × l/µg);

b_i is the intercept of the reference line on the ordinate (its unit depends on the evaluation, e.g. area values).

If data taking recovery into account are required, the mass concentration, ρ_{ic}, of the substance i is calculated using equation (6) after solving equation (1) for the mass concentration, ρ_{ic}.

$$\rho_{ic} = \frac{y_i - b_i}{m_i \times A_i} \qquad ...(6)$$

where

ρ_{ic} is the mass concentration of the substance i in the water sample (corrected by mean recovery), in micrograms per litre;

y_i is the measured value of the substance i in the extract of the water sample (on condition that the same procedure is applied as with the calibration and the sample measurement), (its unit depends on the evaluation, e.g. area value);

m_i is the slope of the calibration curve (**8.1** or **8.4**) of the substance i (its unit depends on the evaluation, e.g. area value × l/µg);

b_i is the intercept of the reference line on the ordinate (its unit depends on the evaluation, e.g. area value);

A_i is the specific mean recovery for the substance i (dimensionless).

9.2.2 *Evaluation using the (re)calibration according to 8.2*

Calculate the mass concentration, ρ_{ig}, of the substance i in the water sample using equation (7), after solving equation (2) for the mass concentration, ρ_{ig}.

$$\rho_{ig} = \frac{y_{ig} - b_{ig}}{m_{ig}} \qquad ...(7)$$

where

ρ_{ig} is the mass concentration of the substance i in the water sample (corrected by recovery), expressed in micrograms per litre;

y_{ig} is the measured value of the substance i in the extract of the water sample (on condition that the same procedure is applied as with the calibration and the sample measurement), (its unit depends on the evaluation, e.g. area value);

m_{ig} is the slope of the calibration curve (**8.2** or **8.4**) of the substance i (its unit depends on the evaluation, e.g. area value × l/µg);

b_{ig} is the intercept of the reference line on the ordinate (its unit depends on the evaluation, e.g. area value).

9.3 Summary of results

When the described procedure is applied, the gas chromatography provides one individual result for each column used. Derive the quantitative final result from these two individual results as follows.

— Take the arithmetic mean, provided the differences between the individual results are less than 10 %, related to the lower result.

— Choose the smaller value in the event of larger differences, provided that the smaller value is not caused by leakage in the gas chromatographic system. The larger value may be the result of peak overlap. Such a result shall be reported as a measured value, obtained from a single separation only.

10 Expression of results

The mass concentrations of the halogenated hydrocarbons of low volatility shall be reported in micrograms per litre:

— at mass concentrations less than 0,01 µg/l, with one significant figure;

— at mass concentrations greater than 0,01 µg/l, with two significant figures.

Table 3 — Characteristic data of the method using hexane as extraction solvent

Compound	l	n	o	ρ	$\bar{x}$	s_r	VC_r	s_R	VC_R	A[b]
			%	ng/l[a]	ng/l[a]	ng/l	%	ng/l	%	%
Matrix drinking water										
Hexachlorobenzene	22	83	4,2	21,8	17,7	3,6	20,5	7,4	42,0	81,0
β-Endosulfan	18	64	6,3	26,5	4,9	1,4	28,4	3,8	78,7	18,4
PCB 180	21	75	5,3	52,1	18,2	4,3	23,6	11,0	60,4	35,0
PCB 180[c]	21	75	5,3	24,0	18,2	4,3	23,6	11,0	60,4	76,0
1, 2, 4, 5-Tetra-chlorobenzene	10	38	10,5	66,7	57,0	6,7	11,8	15,3	26,9	85,5
α-HCH	6	21	19,0	1,3	1,6	0,6	35,6	0,8	49,8	124,4
Dieldrin	16	58	3,5	7,4	6,7	1,3	19,1	2,6	38,4	90,7
p,p'-DDE	17	61	9,8	20,8	11,6	2,0	17,4	6,0	51,9	55,8
p,p'-DDT	11	39	10,3	64,7	35,7	12,1	33,9	22,9	64,1	55,1
PCB 28	5	15	20,0	1,4	1,4	0,6	43,3	0,7	48,7	101,7
Matrix surface water										
Hexachlorobenzene	15	63	0	57,9	48,8	6,6	13,5	16,6	34,1	84,2
β-Endosulfan	14	61	6,6	193,7	21,2	3,6	16,9	14,4	67,9	10,9
PCB 180	15	61	0	217,3	165,2	32,0	20,0	55,6	33,7	76,0
1, 2, 4-Trichlorobenzene	10	44	0	182,2	160,6	18,9	11,8	55,5	34,5	88,1
γ-HCH	15	63	14,3	38,6	37,3	6,5	17,4	14,3	38,4	96,7
Heptachlor	13	50	0	72,9	22,9	4,4	19,3	8,4	36,7	31,4
Dieldrin	14	58	0	30,6	33,3	13,5	40,5	17,2	51,7	108,7
Endrin	14	61	9,8	51,0	50,0	6,8	13,6	11,1	22,3	98,0
o,p'-DDT	9	33	15,2	15,0	17,8	5,9	33,1	6,6	37,0	118,5
PCB 28	13	55	7,3	45,4	41,8	7,3	17,5	14,0	33,5	92,0
PCB 52	13	53	0	74,4	86,9	18,3	21,1	32,7	37,6	116,8
PCB 101	14	49	0	15,2	19,0	6,6	34,9	10,9	57,5	124,9
PCB 138	11	33	9,1	4,3	7,4	1,1	15,0	3,7	49,5	172,1
PCB 153	13	52	5,8	136,9	103,6	16,7	16,1	30,0	28,9	75,7
PCB 194	11	45	0	72,3	56,0	9,1	16,3	21,5	38,3	77,5

l is the number of data sets (i.e. number of laboratories which reported quantitative results)
n is the number of values
o is the percentage of outliers
ρ is the reference concentration
$\bar{x}$ is the mean value, without outliers
s_r is the repeatability standard deviation
VC_r is the repeatability variation coefficient
s_R is the reproducibility standard deviation
VC_R is the reproducibility variation coefficient
A is the recovery (not identical with A_i according to **8.3**)

[a] Value expressed in nanograms per litre.

[b] Related to the reference concentration.

[c] During preparation and partitioning of the intercomparaison sample (total volume 150 litres of drinking water), it was immediately discovered by analysis that losses of PCB 180, caused by absorption, had occurred. Because of this, the reference concentration of 52,1 ng/l was estimated to be 24,0 ng/l. As a consequence, the latter value was taken as the new reference concentration by convention.

11 Performance data

Data from an interlaboratory test organized by Germany with the participation of French, Dutch and British laboratories are shown in Table 3.

12 Test report

The report shall contain the following information:

a) a reference to this International Standard;

b) identity of the original sample;

c) sample pretreatment, if performed;

d) extraction solvent used, procedures used for extraction, concentration, clean-up and separation (by reference to the relevant clauses of this International Standard);

e) evaluation function used according to **9.2**;

f) expression of results according to clause **10**;

g) any deviation from this method and all circumstances which might have affected the results.

Annex A (normative)
Clean-up of the extract by means of pyrogenic copper for the elimination of sulfur

A.1 Preparation of the pyrogenic copper

WARNING — Pyrogenic copper is inflammable in air.

Put 45 g of copper sulfate ($CuSO_4$), 20 ml of hydrochloric acid (2 mol/l) and 480 ml of water in a beaker. Dissolve the copper sulfate.

Put 15 g of coarse powdered zinc in another beaker, of nominal capacity 1 litre. Add 25 ml of water and one drop of wetting agent (based on sodium alkylsulfate), or a spatula tipful of sodium dodecylsulfate.

Mix it into a slurry, with the help of a magnetic stirrer.

While the magnetic bar is running at its highest speed, add slowly, along a glass rod, the copper sulfate solution to the zinc slurry.

When the mixture becomes red and emission of hydrogen occurs, continue the stirring until the emission of gas has considerably diminished. Allow the copper to settle and then decant the water.

Thoroughly wash out the remaining salts using degassed water.

For elimination of the water, use three portions of 250 ml of acetone, shake the beaker, let the copper settle and decant the liquid.

Eliminate the remaining acetone by washing with hexane.

Transfer the hexane/copper suspension into an Erlenmeyer-flask with a ground-glass stopper, close hermetically and store at about 4 °C.

The efficiency decreases during storage after several months. This is indicated by a change of colour.

A.2 Application

Put 1 ml to 2 ml of the sample extract in a centrifuge tube.

Add about 100 mg of copper (see **A.1**), close carefully and mix in an ultrasonic bath for 10 min.

Subsquently, centrifuge for at least 5 min at about 3 500 r/min.

Decant off the extract, wash the copper with hexane and add the washings to the extract.

Annex B (informative)
Examples of gas chromatographic conditions and the corresponding gas chromatographs

Table B.1 — Example of the sequence of elution, relative retention time and retention time, obtained with a non-polar capillary column

Current No.	Compound	Relative retention time	Retention time min
1	1,2,4 Trichlorobenzene	0,293	8,05
2	1,2,3,4-Tetrachlorobenzene	0,543	14,93
3	Pentachlorobenzene	0,679	18,68
4	α-Hexachlorocyclohexane	0,804	22,12
5	β-Hexachlorocyclohexane	0,827	22,74
6	Hexachlorobenzene	0,836	22,99
7	γ-Hexachlorocyclohexane	0,847	23,30
8	δ-Hexachlorocyclohexane	0,861	23,69
9	Pentachloronitrobenzene	0,864	23,76
10	ε-Hexachlorocyclohexane	0,881	24,22
11	Heptachlor	0,956	26,30
12	Aldrin	1,000	27,50
13	Heptachlor-epoxide	1,047	28,80
14	*o,p′*-DDE	1,081	29,74
15	α-Endosulfan	1,086	29,87
16	PCB 101	1,116	30,70
17	Dieldrin	1,119	30,78
18	*p,p′*-DDE	1,124	30,92
19	Endrin	1,137	31,26
20	β-Endosulfan	1,141	31,39
21	*p,p′*-TDE	1,162	31,95
22	*o,p′*-DDT	1,170	32,17
23	PCB 153	1,186	32,62
24	*p,p′*-DDT	1,208	33,22
25	PCB 138	1,213	33,36
26	Methoxychlor	1,271	34,94
27	PCB 180	1,292	35,53

NOTE The gas chromatogram is shown in Figure B.1.

Table B.2 — Example of the sequence of elution, relative retention time and retention time, obtained with a weakly polar capillary column

Current No.	Compound	Relative retention time	Retention time min
1	1, 2, 4 Trichlorobenzene	0,282	7,61
2	1, 2, 3, 4-Tetrachlorobenzene	0,554	14,97
3	Pentachlorobenzene	0,674	18,19
4	Hexachlorobenzene	0,823	22,23
5	α-Hexachlorocyclohexane	0,893	24,10
6	Pentachloronitrobenzene	0,909	24,53
7	γ-Hexachlorocyclohexane	0,949	25,61
8	Heptachlor	0,969	26,15
9	Aldrin	1,000	27,00
10	β-Hexachlorocyclohexane	1,056	28,51
11	ε-Hexachlorocyclohexane	1,060	28,62
12	δ-Hexachlorocyclohexane	1,079	29,12
13	Heptachlor-epoxide	1,079	29,61
14	PCB 101	1,110	29,98
15	*o,p′*-DDE	1,114	30,07
16	α-Endosulfan	1,121	30,27
17	*p,p′*-DDE	1,154	31,16
18	Dieldrin	1,165	31,46
19	Endrin	1,189	32,10
20	*o,p′*-DDT	1,206	32,56
21	PCB 153	1,210	32,67
22	β-Endosulfan		
23	PCB 138	1,249	33,71
24	*p,p′*-TDE		
25	*p,p′*-DDT	1,265	34,15
26	PCB 180	1,324	35,74
27	Methoxychlor	1,340	36,19

NOTE The gas chromatogram is shown in Figure B.2.

Table B.3 — Gas chromatographic conditions

Reference of table	Open tubular column		Carrier gas	Injector	Detector		Oven temperature programme
	Stationary phase Column material Manufacturer Column performance	Length Diameter Film thickness	Gas Pressure at column head Gas flow/linear flow velocity	Temperature Injector mode Split ratio Flow rates a) split vent b) septum purge	Type Operation mode Temperature	Make-up gas Gas ratio Gas flow	
B.1	DB-1 fused silica J & W Scientif. Inc. 2 313 m^{-1}, 170 °C	30 m 0,25 mm 0,25 µm	He 1 bar about 25 cm/s	280 °C split about 10 : 1 a) 50 ml/min b) 5 ml/min	ECD, Ni-63 pulsed 350 °C	Ar/CH_4 95 : 5 (*V*/*V*) 40 ml/min	80 °C for 8 min; 6 °C/min to 250 °C, 20 min isothermal
B.2	DB-1701 fused silica J & W Scientif. Inc. 1 475 m^{-1}, 170 °C	30 m 0,25 mm 0,25 µm	He 1 bar about 25 cm/s	280 °C split about 10 : 1 a) 50 ml/min b) 5 ml/min	ECD, Ni-63 pulsed 350 °C	Ar/CH_4 95 : 5 (*V*/*V*) 40 ml/min	80 °C for 8 min; 6 °C/min to 250 °C, 20 min isothermal

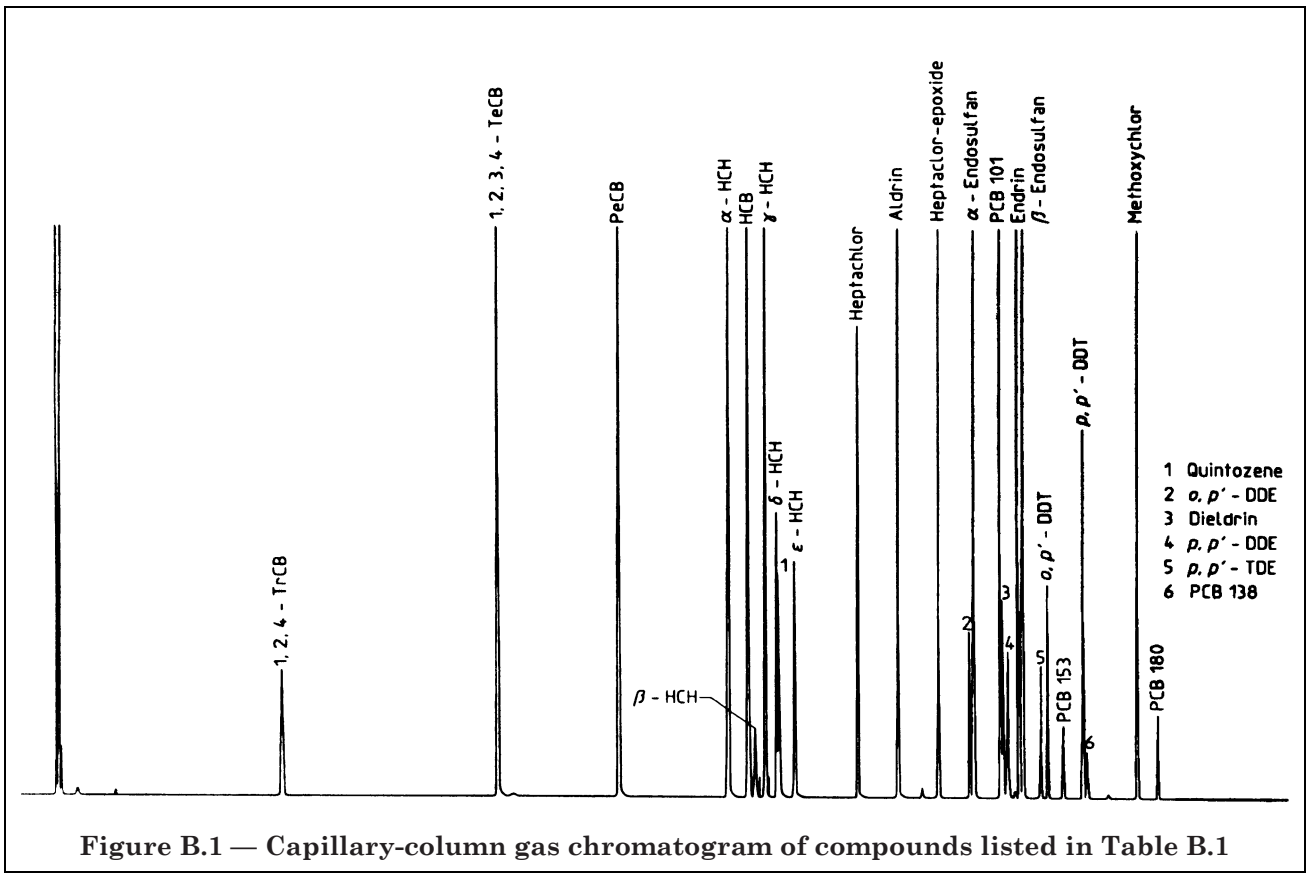


Figure B.1 — Capillary-column gas chromatogram of compounds listed in Table B.1

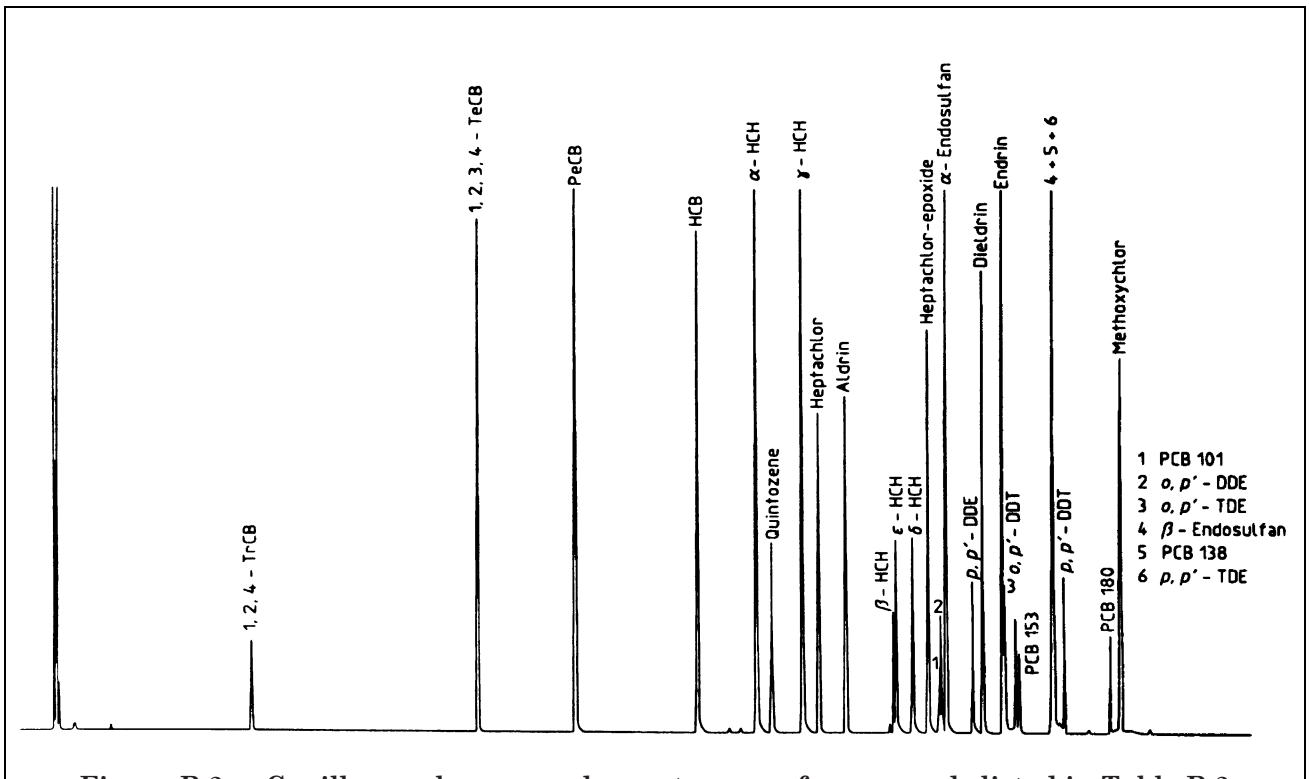


Figure B.2 — Capillary-column gas chromatogram of compounds listed in Table B.2

Annex C (informative)

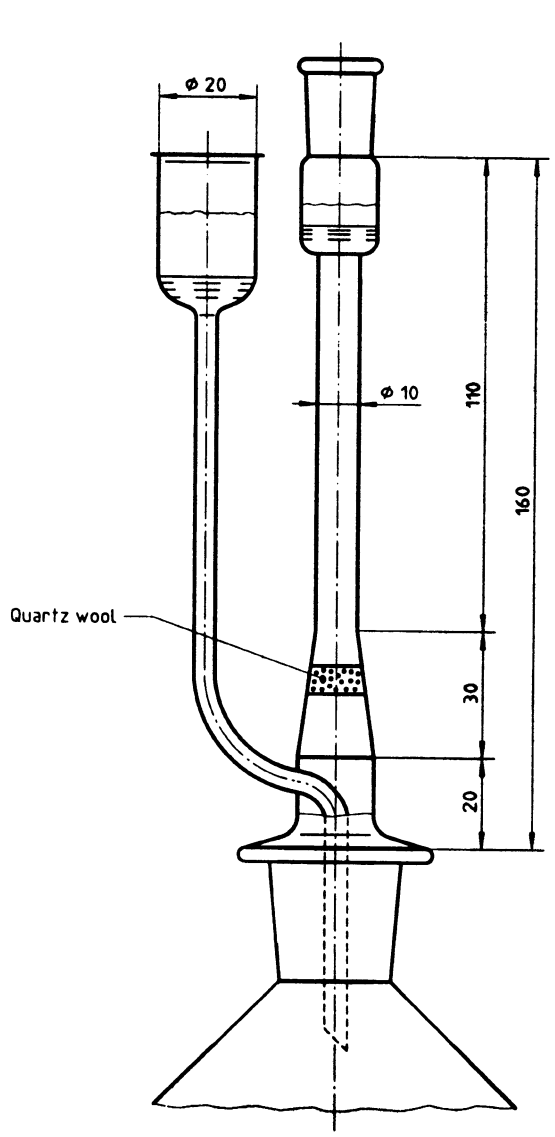


Figure C.1 — Microseparator

Annex D (informative)

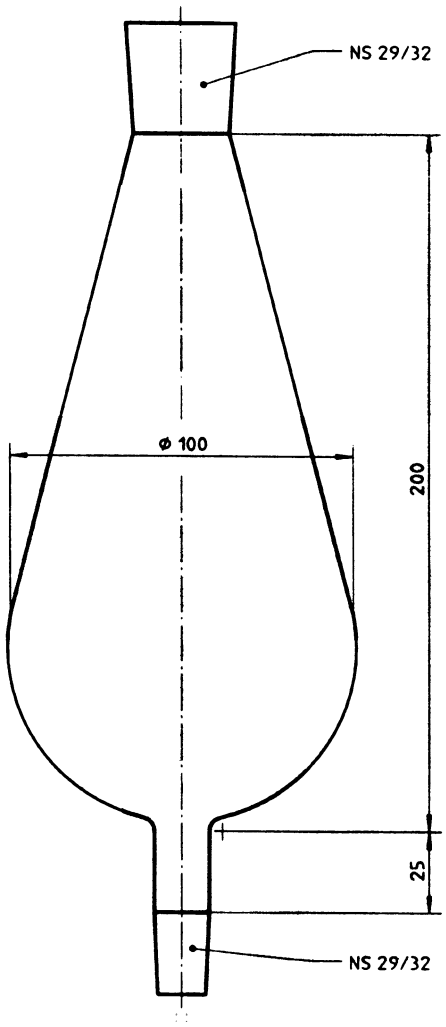


Figure D.1 — Kuderna-Danish evaporation flask

Annex E (informative)

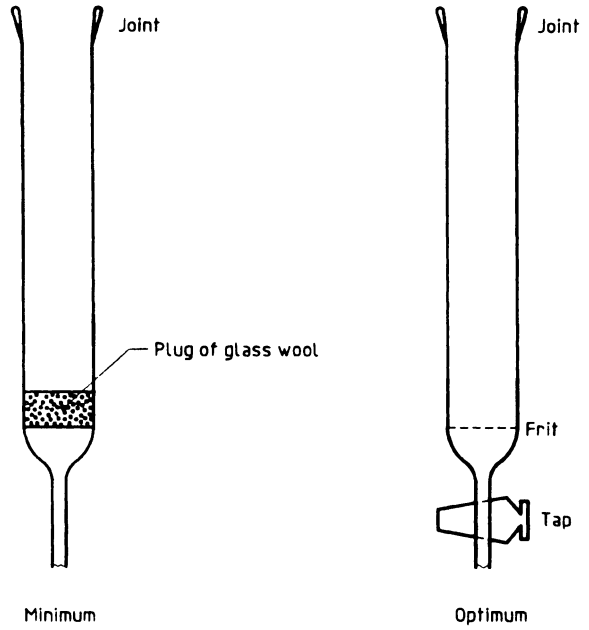


Figure E.1 — Drying and clean-up columns

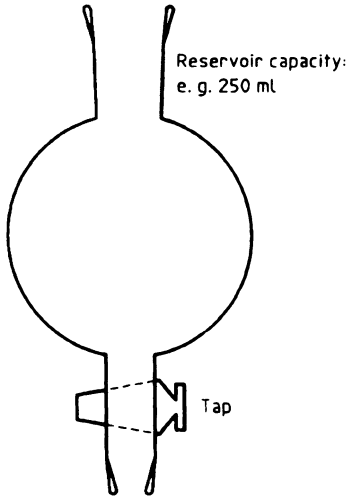
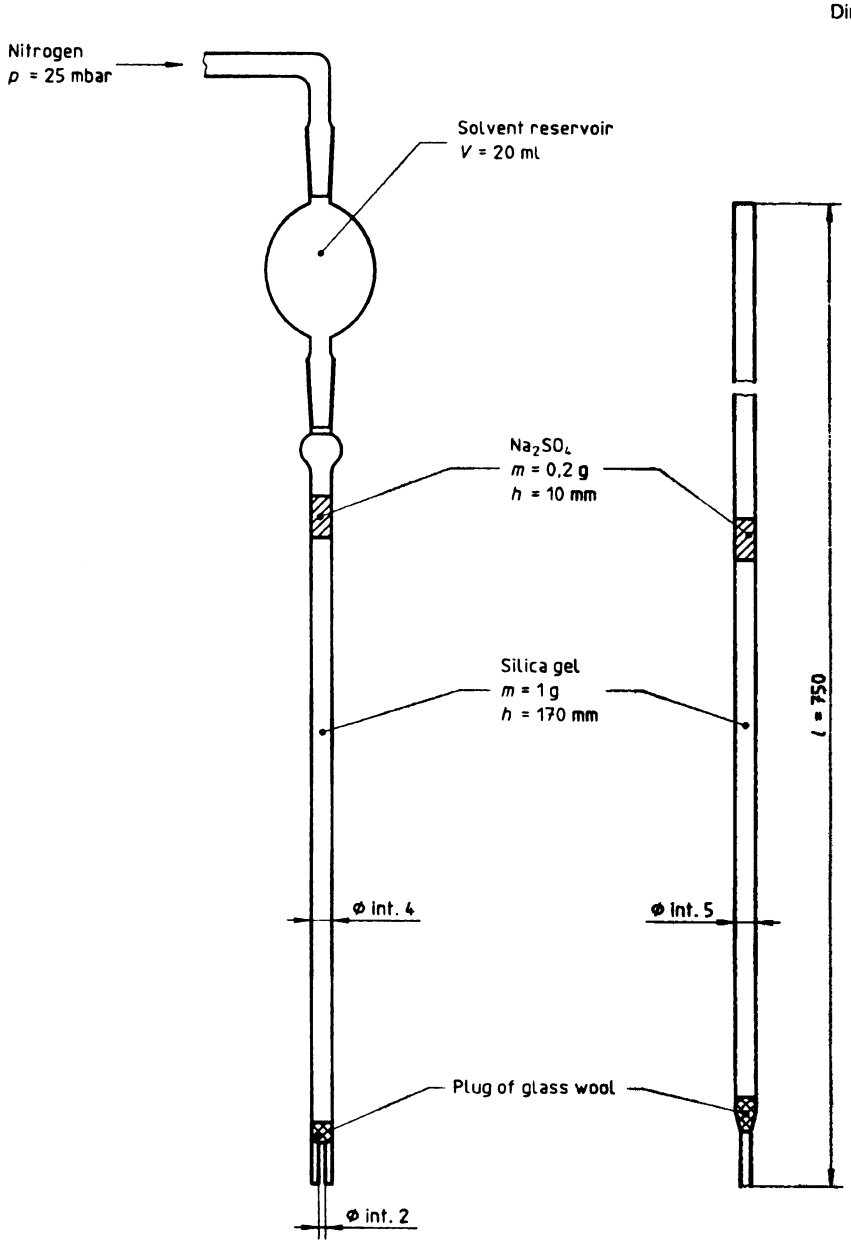


Figure E.2 — Reservoir (to fit top joint of column)

Annex F (informative)



Key

h	Height
m	Mass
p	Pressure
V	Volume

Figure F.1 — Microcolumn for the silica gel clean-up

Annex G (informative)
Clean-up on silica gel macrocolumn

Table G.1 — Sequence of elution of 27 halogenated hydrocarbons using the silica gel clean-up with a macrocolumn

Compound	Recovery[a] of the respective compounds in the various fractions, given in percent of the original amount							
	Fraction 1 Hexane 5 ml	Fraction 2 Hexane 2,5 ml	Fraction 3 Hexane 2,5 ml	Fraction 4 Hexane/ Toluene 95 : 5 8 ml	Fraction 5 Hexane/ Toluene 90 : 10 8 ml	Fraction 6 Hexane/ Toluene 65 : 35 16 ml	Fraction 7 Hexane/ Diethyl-ether 99,5 : 0,5 8 ml	Total
1, 2, 4-TrCB	82 ± 6	1 ± 1						82 ± 7
1, 2, 4, 5-TeCB	85 ± 6	1 ± 1						85 ± 6
PeCB	89 ± 6	2 ± 2						90 ± 6
HCB	97 ± 5	3 ± 3						99 ± 4
PCB 101	[b]	[b]						[b]
PCB 194	101 ± 15	1 ± 1						102 ± 14
PCB 180	103 ± 21	1 ± 1						104 ± 18
PCB 153	101 ± 13	1 ± 1						102 ± 12
PCB 138	[b]	[b]						[b]
PCB 52	99 ± 7	2 ± 2						101 ± 5
PCB 28	97 ± 9							97 ± 9
Aldrin	98 ± 13	2 ± 2						100 ± 11
p,p′-DDE	106 ± 18	3 ± 2						109 ± 17
Heptachlor	65 ± 15	22 ± 25	2 ± 1					88 ± 19
o,p′-DDT	36 ± 20	56 ± 10						92 ± 13
PCNB	9 ± 10	87 ± 11	2 ± 2					97 ± 6
p,p′-DDT		[b]	[b]	[b]				[b]
p,p′-TDE			7 ± 6	106 ± 12	1 ± 1			114 ± 11
α-HCH			2 ± 1	119 ± 12				119 ± 12
γ-HCH				58 ± 20	62 ± 18			121 ± 13
β-HCH				31 ± 16	82 ± 16			115 ± 8
Heptachlor-epoxide				2 ± 2	123 ± 12	2 ± 2		126 ± 13
Endrin						141 ± 7		141 ± 7
Dieldrin						134 ± 11		134 ± 11
α-Endosulfan						[b]		[b]
Methoxychlor						134 ± 3	4 ± 4	137 ± 5
β-Endosulfan						120 ± 6	9 ± 7	129 ± 9

NOTE 1 The recoveries from the extraction step are not taken into consideration.

NOTE 2 All fractions were concentrated with the aid of a rotary evaporator, at a bath temperature of 50 °C and under controlled vacuum. (Fractions 1 to 5 at a pressure of 345 mbar; fraction 6 at the beginning at 345 mbar, until the distillation of hexane was finished; afterwards, the final concentration was achieved at 70 mbar ± 10 mbar; fraction 7 at a pressure of 70 mbar ± 10 mbar.)

[a] Recovery, for the clean-up, including the subsequent step of concentration. Figures given in percent of the original amount of the respective substance (= 100 %) and presented together with the standard deviation, resulting from experiments.

[b] No quantitative data available until now.

Annex H (informative)

The data in Table H.1 were reported from several laboratories in Germany in 1989. The typical mean recoveries are valid for individual analytical conditions.

Table H.1 — Typical mean recoveries, A_i, according to 8.3

No.	Compound	Recoveries, A_i, in laboratories A to I %								
		A	B	C	D	E	F	G	H	I
1	Hexachlorobenzene	75	86	67	69	81	90	102	107	93
2	β-Endosulfan	95	95	81	87	93	97	—	—	—
3	PCB 180	100	100	77	82	89	98	110	80	68
4	1, 2, 4-Trichlorobenzene	45	40	44	48	88	89	—	—	—
5	1, 2, 4, 5-Tetrachlorobenzene	64	—	52	56	78	92	—	—	—
6	Pentachlorobenzene	—	—	—	—	87	94	66	95	89
7	α-HCH	—	—	69	75	—	96	96	118	107
8	β-HCH	—	—	55	63	—	—	—	—	—
9	γ-HCH	85	70	68	74	81	93	96	113	103
10	Pentachloronitrobenzene	—	—	—	—	86	—	—	—	—
11	Heptachlor	—	50	—	—	97	103	75	100	92
12	Aldrin	—	—	—	—	87	—	99	102	110
13	Heptachlor-epoxide	—	—	—	—	87	—	—	—	—
14	α-Endosulfan	—	—	—	—	88	95	110	122	118
15	Dieldrin	90	90	77	86	82	89	92	110	105
16	Endrin	—	100	—	—	82	102	81	82	80
17	*p,p′*-DDE	92	—	75	86	103	—	103	97	90
18	*p,p′*-DDD	—	—	—	—	77	—	142	129	112
19	*o,p′*-DDT	—	92	—	—	92	103	81	98	87
20	*p,p′*-DDT	100	—	76	88	67	104	88	95	83
21	Methoxychlor	—	—	—	—	104	—	23	55	62
22	PCB 28	86	86	—	—	91	98	—	—	—
23	PCB 52	—	86	74	79	94	97	71	94	85
24	PCB 101	—	90	—	—	98	99	72	93	88
25	PCB 138	100	100	76	86	84	99	—	—	—
26	PCB 153	100	100	—	—	85	101	95	86	76
27	PCB 194	—	100	—	—	—	99	—	—	—
28	*o,p′*-DDD	—	—	—	—	—	—	—	—	—

Annex ZA (normative)
Normative references to international publications with their relevant European publications

This European Standard incorporates by dated or undated reference, provisions from other publications. These normative references are cited at the appropriate places in the text and the publications are listed hereafter. For dated references, subsequent amendments to or revisions of any of these publications apply to this European Standard only when incorporated in it by amendment or revision. For undated references the latest edition of the publication referred to applies.

Publication	Year	Title	EN	Year
ISO 5667-1	1980	Water quality — Sampling — Part 1: Guidance on the design of sampling programmes	EN 25667-1	1993
ISO 5667-2	1991	Water quality — Sampling — Part 1: Guidance on sampling techniques	EN 25667-2	1993

List of references

See national foreword.

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