



Standard Test Method for Elemental, Oxidized, Particle-Bound and Total Mercury in Flue Gas Generated from Coal-Fired Stationary Sources (Ontario Hydro Method)¹

This standard is issued under the fixed designation D6784; the number immediately following the designation indicates the year of original adoption or, in the case of revision, the year of last revision. A number in parentheses indicates the year of last reapproval. A superscript epsilon (ϵ) indicates an editorial change since the last revision or reapproval.

1. Scope

1.1 This test method applies to the determination of elemental, oxidized, particle-bound, and total mercury emissions from coal-fired stationary sources.

1.2 This test method is applicable to elemental, oxidized, particle-bound, and total mercury concentrations ranging from approximately 0.5 to 100 $\mu\text{g}/\text{Nm}^3$.

1.3 This test method describes equipment and procedures for obtaining samples from effluent ducts and stacks, equipment and procedures for laboratory analysis, and procedures for calculating results.

1.4 This test method is applicable for sampling elemental, oxidized, and particle-bound mercury in flue gases of coal-fired stationary sources. It may not be suitable at all measurement locations, particularly those with high particulate loadings, as explained in Section 16.

1.5 Method applicability is limited to flue gas stream temperatures within the thermal stability range of the sampling probe and filter components.

1.6 The values stated in SI units are to be regarded as the standard. The values in parentheses are for information only.

1.7 This standard requires users to be familiar with EPA stack-gas sampling procedures as stated in EPA Methods 1–4, Method 5, and Method 17.

1.8 The method requires a high level of experience and quality control both in the field testing and analytical procedures in order to obtain high quality data.

1.9 *This standard does not purport to address all of the safety concerns, if any, associated with its use. It is the responsibility of the user of this standard to establish appropriate safety and health practices and determine the applicability of regulatory limitations prior to use.*

¹ This test method is under the jurisdiction of ASTM Committee D22 on Air Quality and is the direct responsibility of Subcommittee D22.03 on Ambient Atmospheres and Source Emissions.

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2. Referenced Documents

2.1 ASTM Standards:²

- D1193 Specification for Reagent Water
- D1356 Terminology Relating to Sampling and Analysis of Atmospheres
- D3154 Test Method for Average Velocity in a Duct (Pitot Tube Method)
- D3685/D3685M Test Methods for Sampling and Determination of Particulate Matter in Stack Gases
- D3796 Practice for Calibration of Type S Pitot Tubes
- D4840 Guide for Sample Chain-of-Custody Procedures
- D7036 Practice for Competence of Air Emission Testing Bodies
- E2251 Specification for Liquid-in-Glass ASTM Thermometers with Low-Hazard Precision Liquids

2.2 Other Standards:³

- EPA Method 1 Sample and Velocity Traverses for Stationary Sources
- EPA Method 2 Determination of Stack Gas Velocity and Volumetric Flow Rate (Type S Pitot Tube)
- EPA Method 3 Gas Analysis for the Determination of Dry Molecular Weight
- EPA Method 4 Determination of Moisture Content in Stack Gases
- EPA Method 5 Determination of Particulate Emissions from Stationary Sources
- EPA Method 12 Determination of Inorganic Lead Emissions from Stationary Sources
- EPA Method 17 Determination of Particulate Emissions from Stationary Sources (In-Stack Filtration Method)
- EPA Method 29 Determination of Metals Emissions from Stationary Sources

² For referenced ASTM standards, visit the ASTM website, www.astm.org, or contact ASTM Customer Service at service@astm.org. For *Annual Book of ASTM Standards* volume information, refer to the standard's Document Summary page on the ASTM website.

³ EPA Methods 1 – 29 available from the U.S. Environmental Protection Agency's Emission Measurement Technical Information Center or Code of Federal Regulations (40 CFR Part 60, Appendix A), Method 101A in 40 CFR Part 61, Appendix B, Method 301 in 40 CFR 63 Appendix A40 CFR Part 61, Appendix B.

EPA Method 101A Determination of Particle-Bound and Gaseous Mercury Emissions from Sewage Sludge Incinerators

EPA Method 301 Field Validation of Pollutant Measurement Methods from Various Waste Media

EPA SW 846 7470A Mercury in Liquid Waste—Manual Cold Vapor Technique

EPA Water and Waste 600/4-79-020 Methods for Chemical Analysis of Water and Wastes

3. Terminology

3.1 Definitions other than those given below in 3.2 and 3.3 are listed in Terminology **D1356**.

3.2 *Definitions of Terms Specific to This Standard:*

3.2.1 *elemental mercury*—mercury in its zero oxidation state, Hg⁰.

3.2.2 *elemental mercury catch*—mercury collected in the acidified hydrogen peroxide (HNO₃–H₂O₂) and potassium permanganate (H₂SO₄–KMnO₄) impinger solutions employed in this test method. This is gaseous Hg⁰.

3.2.3 *front half of the sampling train*—all mercury collected on and upstream of the sample filter.

3.2.4 *impinger train*—setup including only the impingers and connectors.

3.2.5 *method detection limit*—the minimum concentration of an analyte, when processed through the complete method, produces a signal with a 99 % probability that is different from the blank, based on a standard deviation of greater than seven replicate measurements (see Terminology **D1356**).

3.2.6 *oxidized mercury*—mercury in its mercurous or mercuric oxidation states: Hg₂²⁺ and Hg²⁺, respectively.

3.2.7 *oxidized mercury catch*—mercury collected in the aqueous potassium chloride (KCl) impinger solution employed in this test method. This is gaseous Hg²⁺.

3.2.8 *particle-bound mercury catch*—mercury associated with the particulate matter collected in the front half of the sampling train.

3.2.9 *sample train*—complete setup including nozzle, probe, probe liner, filter, filter holder, impingers, and connectors.

3.2.10 *total mercury*—all mercury (solid-bound, liquid, or gaseous) however generated or entrained in the flue gas stream (that is, summation of elemental, oxidized, and particle-bound mercury).

3.3 *Symbols:*

A = cross-sectional area of stack, m² (ft²)

B_{ws} = water vapor in the gas stream, proportion by volume

ΔH = average pressure differential across the orifice meter, kPa (in. H₂O)

Hg_{ash} = concentration of mercury in sample filter ash, µg/g

Hg^{ip} = concentration of particle-bound mercury, µg/Nm³

Hg^0 = concentration of elemental mercury, µg/Nm³

Hg^{2+} = concentration of oxidized mercury, µg/Nm³

IR = instrument reading from mercury analyzer, µg/L

L_p = leakage rate observed during the post test leak check, m³/min (cfm)

L_a = maximum acceptable leakage rate

M_s = molecular weight of stack gas, wet basis g/g-mole (lb/Lb-mole)

M_w = molecular weight of water, 18.0 g/g-mole (18.0 lb/Lb-mole)

N = Normal conditions, defined as 0°C and 101.3 kPa, (In the U.S. standard conditions 32°F and 1 atmosphere)

P_{bar} = barometric pressure at the sampling site, kPa (in. Hg)

P_s = absolute stack gas pressure, kPa (in. Hg)

P_{std} = standard absolute pressure, 101.3 kPa (29.92 in. Hg)

R = ideal gas constant, 0.008314 kPa·m³/K·g-mole (21.85 in. Hg·ft³/°R·lb-mole)

T_m = absolute average dry gas meter temperature, K (°R)

T_s = absolute stack temperature, K (°R)

T_{std} = standard absolute temperature, 293 K (528°R)

V_D = total digested volume, mL

V_m = volume of gas sample as measured by dry gas meter, m³ (dscf)

$V_{m(std)}$ = volume of gas sample measured by the dry gas meter, corrected to standard conditions, Nm³ (dscf)

$V_{w(std)}$ = volume of water vapor in the gas sample, corrected to standard conditions, m³ (scf)

W_{ash} = total mass of ash on sample filter, g

W_{lc} = total weight of liquid collected in impingers and silica gel, g (lb)

Y = dry gas meter calibration factor

θ = total sampling time, min

θ_1 = sampling time interval, from the beginning of a run until the first component change, min

4. Summary of Test Method

4.1 A sample is withdrawn from the flue gas stream isokinetically through a probe/filter system, maintained at 120°C or the flue gas temperature, whichever is greater, followed by a series of impingers in an ice bath. Particle-bound mercury is collected in the front half of the sampling train. Oxidized mercury is collected in impingers containing a chilled aqueous potassium chloride solution. Elemental mercury is collected in subsequent impingers (one impinger containing a chilled aqueous acidic solution of hydrogen peroxide and three impingers containing chilled aqueous acidic solutions of potassium permanganate). Samples are recovered, digested, and then analyzed for mercury using cold-vapor atomic absorption (CVAAS) or fluorescence spectroscopy (CVAFS). To achieve the precision specified in this test method, it is necessary that quality control and quality assurance procedures associated with each step of the method be scrupulously performed. Successful performance of the method by air emission testing bodies is best achieved by following the Practice **D7036**.

5. Significance and Use

5.1 The measurement of particle-bound, oxidized, elemental, and total mercury in stationary-source flue gases provides data that can be used for emissions assessments and reporting, the certification of continuous mercury monitoring systems, regulatory compliance determinations and research programs associated with dispersion modeling, deposition

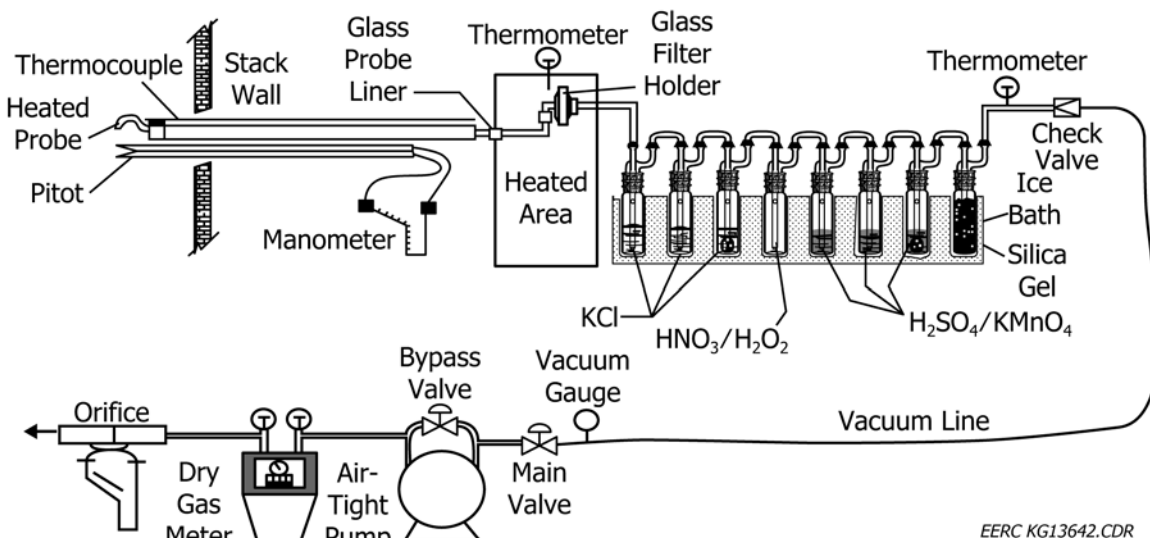


FIG. 1 Schematic of Mercury-Sampling Train in the Method 5 Configuration

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evaluation, human health and environmental impact assessments. Particle-bound, oxidized, and elemental mercury measurements before and after control devices may be necessary for optimizing and evaluating the mercury removal efficiency of emission control technologies.

5.2 This test method was developed for the measurement of mercury in coal-fired power plants and has been extensively validated for that application. With additional procedures given in this standard, it is also applicable to sources having a flue gas composition with high levels of hydrochloric acid, and low levels of sulfur dioxide.

6. Interferences

6.1 Chlorine and particulate matter will interfere in speciating flue gas samples for oxidized and elemental mercury concentrations. These biases are addressed further in Section 16 of this test method.

7. Apparatus

7.1 *Sampling Train*—Similar to Test Methods D3685/D3685M, EPA Method 5/EPA Method 17 and EPA Method 29 trains, as illustrated in Fig. 1 and Fig. 2.

NOTE 1—It is recommended that an in-stack filter method (Method 1, Figure 2) be used if possible. The requirement of the method, that the filter be maintained at the temperature of the flue gas, is ensured in this configuration. In addition, the instack filter method has the added advantage that, only a small portion of the probe/nozzle collects ash that needs to be brushed onto the filter. Method 5 procedures must be used when the temperature of the flue gas is below the water dew point (wet stack) In this case an out-of-stack filter must be used and maintained at a temperature of 120°C.

NOTE 2—If sampling is conducted in a wet stack where water droplets are present, and the nozzle is positioned into the flow, water droplets will be collected and mercury contained in the droplets will be measured. When water droplets are present, the isokinetic sampling rate and percent isokinetic must be calculated accordingly.

7.1.1 *Probe Nozzle (Probe Tip)*—Glass nozzles are required unless alternate nozzles are constructed of materials that are free from contamination and will not interact with the sample.

Probe fittings constructed of polytetrafluoroethylene (PTFE), polypropylene, etc., are required instead of metal fittings to prevent contamination.

7.1.2 *Probe Liner*—If the sample train is to be in EPA Method 5 configuration (out-of-stack filtration), the probe liner must be constructed of quartz or borosilicate glass. If an EPA Method 17 (in-stack filtration) sampling configuration is used, the probe/probe liner may be constructed of borosilicate glass, quartz or, depending on the flue gas temperature, PTFE.

7.1.3 *Pitot Tube*, Type S pitot tube. Refer to Section 2.2 of EPA Method 2 for a description.

7.1.4 *Differential Pressure Gages*, inclined manometers or equivalent devices. Refer to Section 2.1 of EPA Method 2 for a description.

7.1.5 *Filter Holder*, constructed of borosilicate glass or PTFE-coated stainless steel with a PTFE filter support or other nonmetallic, non-contaminating support. Do not use a glass frit or stainless steel wire screen. A silicone rubber or PTFE gasket, designed to provide a positive seal against leakage from outside or around the filter, may be used.

7.1.6 *Connecting Umbilical Tube*, heated PTFE tubing. This tube must be heated to a minimum of 120°C to help prevent water and acid condensation. (The umbilical tube is defined as any tubing longer than 0.5 m that connects the filter holder to the impinger train).

7.1.7 *Probe and Filter Heating System:*

7.1.7.1 *EPA Method 5 Configuration*—For EPA Method 5 configuration, the temperature of the flue gas, sample probe, and the exit of the sample filter must be monitored using temperature sensors capable of measuring temperature to within 3°C (5.4°F). The heating system must be capable of maintaining the sample gas temperature of the probe and exit of the sample filter to within ±15°C (±27°F) of the flue gas temperature. Regardless of the flue gas temperature, to prevent water and acid condensation, the probe temperature, sample filter exit gas temperature, or the temperature of the connecting umbilical cord shall at no time be less than 120°C.

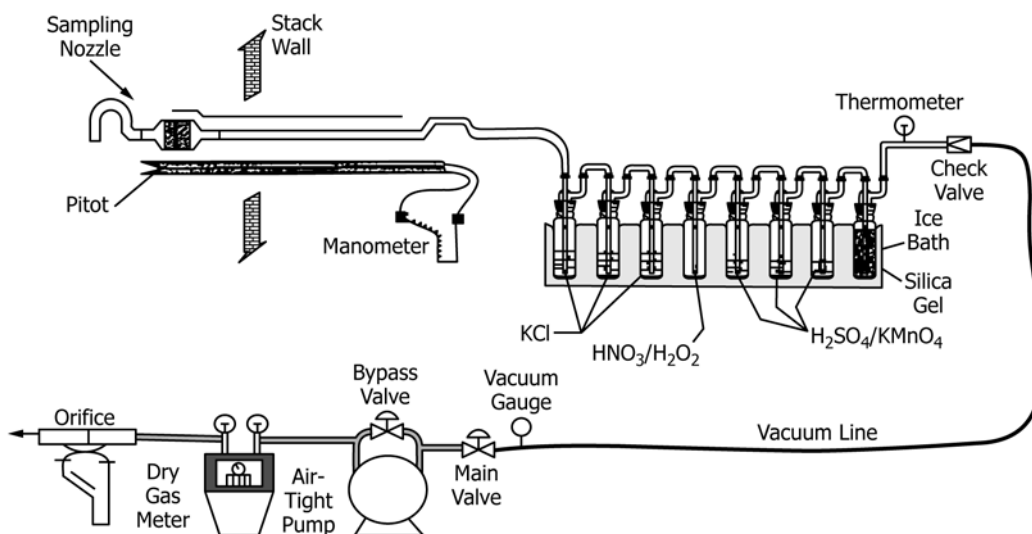


FIG. 2 Schematic of Mercury-Sampling Train in the Method 17 Configuration

7.1.7.2 *EPA Method 17 Configuration*—For EPA Method 17 configuration, the sample filter is located in the duct and, therefore, naturally maintained at the flue gas temperature. The heating system is only required to maintain the probe and connecting umbilical cord to at least 120°C. If the flue gas temperature is less than 120°C, then EPA Method 5 configuration must be used.

7.1.8 *Condensing/Absorbing System*, consists of eight impingers immersed in an ice bath and connected in series with leak-free ground glass fittings or other non-contaminating leak-free fittings. (At no time is silicon grease or other greases to be used for this test method). The first, second, fourth, fifth, sixth, and eighth impingers are of the Greenburg-Smith design modified by replacing the standard tip with a 1.3-cm (0.5-in.)-ID straight glass tube extending to about 1.3 cm (0.5 in.) from the bottom of the flask. The third and seventh impingers are also Greenburg-Smith design, but with the standard tip including the glass impinging plate. The first, second, and third impingers contain aqueous 1 N potassium chloride (KCl) solution. The fourth impinger contains an aqueous solution of 5 % v/v nitric acid (HNO_3) and 10 % v/v hydrogen peroxide (H_2O_2). The fifth, sixth, and seventh impingers contain an aqueous solution of 4 % w/v potassium permanganate ($KMnO_4$) and 10 % v/v sulfuric acid (H_2SO_4). The last impinger contains silica gel or an equivalent desiccant. Refer to [Note 4](#).

NOTE 3—When flue gas streams are sampled with high moisture content (>20 %), additional steps must be taken to eliminate carryover of impinger contents from one sample type to the next. These steps must include use of oversized impinger(s) or use of an empty impinger between the KCl and $HNO_3-H_2O_2$. If a dry impinger is used, it must be rinsed as discussed in [13.2](#) of this test method and the rinse added to the preceding impinger.

7.1.9 *Metering System*, vacuum gage, leak-free pump, thermometers capable of measuring temperature to within 3°C (5.4°F), and a dry gas meter or controlled orifice capable of measuring volume to within 2 %.

7.1.10 *Barometer*, capable of measuring atmospheric pressure to within 0.33 kPa (0.1 in. Hg). In many cases, the barometric reading may be obtained from a nearby National

Weather Service station, in which case, the station value (which is the absolute barometric pressure) shall be requested. An adjustment for elevation differences between the weather station and sampling point shall be applied at a rate of negative 0.33 kPa (0.1 in. Hg) per 30 m (100 ft) elevation increase or vice versa for elevation decrease.

7.1.11 *Thermometers*, Precision digital thermometers based on resistance temperature detectors (RTDs), thermistors, thermocouples, or organic liquid-in-glass thermometers (such as Thermometer S18C in Practice [E2251](#)) meeting the requirements of specific applications in this test method may be used.

7.1.12 *Gas Density Determination Equipment*, temperature sensor and pressure gage, as described in Section 2.3 and 2.4 of EPA Method 2. The temperature sensor shall, preferably, be permanently attached to the pitot tube or sampling probe in a fixed configuration, such that the sensor tip extends beyond the leading edge of the probe sheath and does not touch any metal. Alternative temperature sensor configurations are described in Section 2.1.10 of EPA Method 5. If necessary, a gas analyzer can be used to determine dry molecule weight of the gas (refer to EPA Method 3).

7.2 *Digestion Apparatus:*

7.2.1 *Dry Block Heater or Hot Water Bath*, a heater capable of maintaining a temperature of 95°C is required for digestion of samples, similar to that described in EPA SW 846 Method 7470A.

7.2.2 *Ice Bath.*

7.2.3 *Digestion Flasks*—Use 50- to 70-mL glass tubes or flasks with screw caps that will fit a dry block heater. For a water bath, 300-mL biological oxygen demand glass bottles for SW 846 Method 7470A are to be used. In addition, borosilicate glass test tubes, 35- to 50-mL volume, with rack are needed.

7.2.4 *Microwave or Convection Oven and PTFE Digestion Vessels*, 120 mL, or equivalent digestion vessels with caps equipped with pressure relief valves for the dissolution of ash, along with a capping station or the equivalent to seal the digestion vessel caps. Use a vented microwave or convection

oven for heating. In addition, polymethylpentene (PMP) or equivalent volumetric flasks are recommended for the digested ash solutions.

7.3 Analytical Equipment:

7.3.1 *Mercury Analyzer*, dedicated mercury analyzer or equivalent apparatus for the analysis of mercury via CVAAS. Alternatively, CVAFS may be used. CVAAS is a method based on the absorption of radiation at 253.7 nm by mercury vapor. The mercury is reduced to the elemental state and aerated from solution in a closed system. The mercury vapor passes through a cell positioned in the light path of an atomic absorption spectrometer. Absorbency is measured as a function of mercury concentration. A soda-lime trap and a magnesium perchlorate trap must be used to precondition the gas before it enters the absorption cell.

7.3.2 *Pipettors*—All analysis should be performed with pipettors having accuracy to be within $\pm 0.5\%$ of the true value, and precision $\leq 0.5\%$. A repeater pipettor is recommended to reduce the time required for sample preparation and analysis. Air displacement pipettors are not recommended.

7.3.3 *Transfer pipets*, low-density polyethylene disposable transfer pipets.

7.3.4 *Balance*, analytical grade, capable of weighing the filter and sample beakers to within 0.1 g.

7.4 *Ancillary equipment*, depending upon the application, other flue gas parameters may need to be obtained to convert the mercury measurements into appropriate units. This equipment may include sampling equipment and O₂ or CO₂ analyzers.

7.5 *Spare Parts*—Enough sampling equipment must be brought to the site so that common spare parts are available. Arrangements should be made so that, if necessary, parts can also be shipped next-day to the site.

8. Reagents and Materials

8.1 *Purity of Reagents*—Reagent-grade chemicals shall be used in all tests. Unless otherwise indicated, it is intended that all reagents conform to the specifications of the Committee on Analytical Reagents of the American Chemical Society, where such specifications are available.⁴ Other grades may be used, provided it is first ascertained that the reagent is of sufficiently high purity to permit its use without lessening the accuracy of the determination.

8.2 *Purity of Water*—Unless otherwise indicated, references to water shall be understood to mean reagent water as defined by Type II in Specification **D1193**.

8.3 Reagents:

8.3.1 *Boric Acid (H₃BO₃)*, purified reagent grade.

8.3.2 *Hydrochloric Acid (HCl)*, trace metal-grade concentrated hydrochloric acid, with a specific gravity of 1.18.

8.3.3 *Hydrofluoric Acid (HF)*, concentrated hydrofluoric acid, 48 to 50 %.

8.3.4 *Hydrogen Peroxide (H₂O₂)*, 30 %*v/v* hydrogen peroxide.

8.3.5 *Hydroxylamine Sulfate ((NH₂OH)₂ · H₂SO₄)*, solid.

8.3.6 *Hydroxylamine Hydrochloride (NH₂OH)₂ · HCl*, 10 % solution.

8.3.7 *Sodium Chloride (NaCl)*, solid.

8.3.8 *Mercury Standard Solution*, a certified (1000 µg/mL) mercury standard.

8.3.9 *Nitric Acid (HNO₃)*, trace metal-grade concentrated nitric acid with a specific gravity of 1.42. 20 %*v/v* nitric acid.

8.3.10 *Potassium Chloride (KCl)*, solid.

8.3.11 *Potassium Dichromate (K₂Cr₂O₇)*, solid.

8.3.12 *Potassium Perchlorate (KClO₄)*, solid.

8.3.13 *Potassium Permanganate (KMnO₄)*, solid.

8.3.14 *Potassium Persulfate (K₂S₂O₈)*, solid.

8.3.15 *Soda Lime (Ca(OH)₂, NaOH, KOH)*, solid.

8.3.16 *Sodium Thiosulfate (Na₂S₂O₃ · 5H₂O)* (for high chloride applications).

8.3.17 *Stannous Chloride (SnCl₂ · 2H₂O)*, solid.

8.3.18 *Sulfuric Acid (H₂SO₄)*, trace metal-grade concentrated sulfuric acid, with a specific gravity of 1.84.

8.3.19 *Tin (Sn) Mossy*.

8.4 Materials:

8.4.1 *Indicating Silica Gel*, with a size of 6-16 mesh.

8.4.2 *Crushed or Cubed Ice*.

8.4.3 *Sample Filters*, quartz fiber filters, without organic binders, exhibiting at least 99.95 % efficiency ($< 0.05\%$ penetration) for 0.3-µm dioctyl phthalate smoke particles and containing less than 0.2 µg/m² of mercury. Test data provided by filter manufacturers and suppliers stating filter efficiency and mercury content are acceptable. Filter material must be unreactive to sulfur dioxide (SO₂) or sulfur trioxide (SO₃).⁵

8.4.4 *Filter Papers*, for filtration of digested samples. The filter paper must have a particle retention of $> 20\ \mu\text{m}$ and filtration speed of $> 12\ \text{s}$.

8.4.5 *Nitrogen Gas (N₂)*, carrier gas of at least 99.998 % purity. Alternatively, argon gas may be used.

8.4.6 *Soda Lime*, indicating 4- to 8-mesh absorbent for trapping carbon dioxide.

8.4.7 *Sample Containers*, glass or PTFE with PTFE-lined lids.

NOTE 4—It is recommended that glass amber bottles be used to prevent possible deterioration by ultraviolet (UV) light.

8.5 Sampling Reagents:

8.5.1 *KCl Absorbing Solution (1 mol/L)*—Dissolve 74.56 g of KCl in 500 mL of reagent water in a 1000-mL volumetric flask, swirl to mix, and dilute to volume with water. Mix well. A new batch of solution must be made prior to each field test.

NOTE 5—For applications with High Chloride Applications: KCl Absorbing Solution spiked with Sodium Thiosulfate (1mol/l KCl, 0.5 %*v/v* Na₂S₂O₃ · 5H₂O) – 5 g Na₂S₂O₃ · 5H₂O is dissolved in 1 litre of 1 N KCl solution. This solution is used to charge each impinger (100 ml per impinger). This solution should be made daily.

⁴ “Reagent Chemicals, American Chemical Society Specifications,” Am. Chemical Soc., Washington, DC. For suggestions on the testing of reagents not listed by the American Chemical Society, see “Reagent Chemicals and Standards,” by Joseph Rosin, D. Van Nostrand Co., Inc., New York, NY, and the “United States Pharmacopeia.”

⁵ Felix, L.G.; Clinard, G.I.; Lacey, G.E.; McCain, J.D. “Inertial Cascade Impactor Substrate Media for Flue Gas Sampling,” U.S. Environmental Protection Agency, Research Triangle Park, NC 27711, Publication No. EPA-600/7-77-060; June 1977, p. 83.

8.5.2 *HNO₃–H₂O₂ Absorbing Solution (5 %*v/v* HNO₃, 10 %*v/v* H₂O₂)*—Add slowly, with stirring, 50 mL of concentrated HNO₃ to a 1000-mL volumetric flask containing approximately 500 mL of water, and then add carefully, with stirring, 333 mL of 30 %*v/v* H₂O₂. Dilute to volume with water. Mix well. A new batch of solution must be made prior to each field test.

8.5.3 *H₂SO₄–KMnO₄ Absorbing Solution (4 %*w/v* KMnO₄, 10 %*v/v* H₂SO₄)*—Mix carefully, with stirring, 100 mL of concentrated H₂SO₄ into approximately 800 mL of water. When mixing, be sure to follow standard acid to water addition procedures and safety precautions associated with strong acids. Then add water, with stirring, to make 1 L. This solution is 10 %*v/v* H₂SO₄. Dissolve, with stirring, 40 g of KMnO₄ into 10 %*v/v* H₂SO₄, and add 10 %*v/v* H₂SO₃, with stirring, to make 1 L. To prevent autocatalytic decomposition of the permanganate solution, filter the solution through filter paper. (**Warning**—See 9.1.1.) H₂SO₄–KMnO₄ absorbing solution must be made daily.

8.5.4 *Saturated Potassium Permanganate Solution (5 %*w/v*)*—Mix 5 g KMnO₄ into water, dilute to 100 mL, and stir vigorously.

8.6 Rinse Solutions for Sample Train:

8.6.1 *0.1 N HNO₃ Solution*—A certified reagent grade 0.1 N HNO₃ solution can be purchased directly or can be made by slowly adding 12.5 mL of concentrated HNO₃ to a 2000-mL volumetric flask containing approximately 500 mL of water, then diluting with water to volume.

8.6.2 *10 %*w/v* HNO₃ Solution*—Mix carefully, with stirring, 100 mL of concentrated HNO₃ into approximately 800 mL of water. When mixing, be sure to follow standard acid to water addition procedures and safety precautions associated with strong acids. Then add water, with stirring, to make 1 L.

8.6.3 *10 %*w/v* Hydroxylamine Solution*—Add 100 g hydroxylamine sulfate and 100 g sodium chloride to a 1000-mL volumetric flask containing approximately 500 mL of water. After the hydroxylamine sulfate and sodium chloride has been dissolved, dilute with water to volume. As an alternative a 10 % hydroxylamine hydrochloride solution can be used in all cases as a replacement for the hydroxylamine sulfate/sodium chloride solution.

8.7 Sample Digestion Reagents:

8.7.1 *Boric Acid Solution (4 %*w/v*)*—Dissolve 4 g H₃BO₃ in water, and dilute to 100 mL.

8.7.2 *Aqua Regia (HCl:HNO₃ 3:1)*—Add 3 parts concentrated HCl to 1 part concentrated HNO₃. Note that this should be made up in advance and allowed to form a dark orange color. This mixture should be loosely capped, as pressure will build as gases form.

8.7.3 *Saturated Potassium Permanganate Solution (5 %*w/v*)*—Mix 5 g KMnO₄ into water, dilute to 100 mL, and stir vigorously.

8.7.4 *Potassium Persulfate Solution (5 %*w/v*)*—Dissolve 5 g K₂S₂O₈ in water, and dilute to 100 mL.

8.7.5 *Potassium Dichromate Solution (5 %*w/v*)*—Dissolve 5 g K₂Cr₂O₇ in water, and dilute to 100 mL.

8.8 Analytical Reagents:

8.8.1 *Hydrochloric Acid Solution (10 %*v/v*)*—Add 100 mL concentrated HCl to water, and dilute to 1 L. Be sure to follow all safety precautions for using strong acids.

8.8.2 *Stannous Chloride Solution (10 %*w/v*)*—Dissolve 100 g in 10 %*v/v* HCl, and dilute with 10 %*v/v* HCl to 1 L. Difficulty in dissolving the stannous chloride can be overcome by dissolving in a more concentrated HCl solution (such as 100 mL of 50 %*v/v* HCl) and diluting to 1 L with water. Note that care must be taken when adding water to a strong acid solution. Add a lump of mossy tin (~0.5 g) to this solution.

8.9 Mercury Standards:

8.9.1 *10 µg/L Hg Stock Solution*—Dilute 1 mL of 1000 µg/L Hg standard solution to 100 mL with 10 %*v/v* HCl.

8.9.2 *100 µg/L Hg Stock Solution*—Dilute 1 mL of 10 µg/L Hg stock solution to 100 mL with 10 %*v/v* HCl. This solution and the Working Hg Standards described below may change concentration with time. As a minimum, stock solutions should be prepared weekly, and stored in glass or PTFE bottles.

8.9.3 *Working Hg Standards*—Prepare all working standards by digesting along with the samples. Prepare digested standards of 0.25, 0.5, 1.0, 2.5, 5.0, 7.5, and 10.0 µg/L as described in 13.4.1.1.

8.9.4 *Quality Control Standard (QC)*—A quality control standard is prepared from a separate Hg standard solution. The QC standard should be prepared at a concentration of approximately one-half the calibration range. It is recommended to prepare a QC standard at a concentration of 5.0 µg/L in the same manner as the 5.0 µg/L standard described in 8.9.3.

8.10 *Glassware Cleaning Reagents*—Prior to any fieldwork, all glassware must be cleaned in accordance with the guidelines outlined in EPA Method 29, Section 8.1.1 if the stated precision of this test method is to be met. This procedure requires that the sampling train glassware first be rinsed with hot tap water and then washed in hot soapy water. Then, rinse the glassware three times with tap water, followed by three additional rinses with distilled water. Soak all glassware in a 20 %*v/v* nitric acid solution for a minimum of 4 hours. Rinse three times with distilled water, and rinse a final time with acetone. Allow the glassware to air dry and cover all glassware openings where contamination can occur until the sampling train is assembled for sampling.

NOTE 6—There are two ways to ensure clean glassware. The first is to bring enough glassware into the field to construct all needed sampling trains. The second, is to clean the glassware in the field. This requires a large enough space to soak the glassware. In addition, depending on the scope of the sampling program, an extra person may be required on site.

9. Hazards

9.1 Warning:

9.1.1 Pressure may build up in the solution storage bottle because of a potential reaction between potassium permanganate and acid. Therefore, these bottles should not be fully filled and should be vented to relieve excess pressure and prevent explosion. Venting must be in a manner that will not allow contamination of the solution.

9.1.2 Hazards to personnel exist in the operation of the cold-vapor atomic absorption spectrophotometer. Refer to the manufacturer's instruction manual before operating the instrument.

9.1.3 Sample digestion with hot concentrated acids creates a safety problem. Observe appropriate laboratory procedures for working with concentrated acids. Hydrofluoric acid used in the sample digestion procedures is highly corrosive and is very toxic by inhalation or contact with the skin. Avoid exposure by contact with the skin or eyes, or by inhalation of HF vapor. It is essential to use suitable personal protective equipment, including impermeable gloves and eye protection when working with HF. Use a fume hood when working with concentrated HF and when carry out open-vessel dissolution with HF.

9.1.4 Mercury standards at high concentrations (1000 µg/mL) can cause skin irritation, serious eye damage and may damage fertility of an unborn child. Suitable safety equipment (gloves, goggles, etc.) should be used when working with standards and samples containing mercury, or where exposure to mercury vapors is a concern.

9.1.5 Acetone is hazardous in case of skin contact (irritant, of eye contact (irritant), of ingestion, and of inhalation. Acetone is highly flammable in the presence of open flames or sparks.

9.2 Precaution:

9.2.1 The determination of microquantities of mercury species requires meticulous attention to detail. Good precision is generally unattainable without a high level of experience with stack-sampling procedures. Precision may be improved by knowledge of, and close adherence to, the suggestions that follow.

9.2.1.1 All glassware used in the method must be cleaned thoroughly prior to use in the field, as described in 8.10 of this test method.

9.2.1.2 Use the same reagents and solutions in the same quantities for a group of determinations and the corresponding solution blank. When a new reagent is prepared or a new stock of filters is used, a new blank must taken and analyzed.

10. Sampling

10.1 Preparation for Test:

10.1.1 *Quality Assurance Plan*—Develop a quality assurance plan (QAP) prior to conducting the tests. The basic elements of the QAP are sections that describe: (1) Purpose of the project, (2) Test methodologies, (3) Project organization, (4) Description of test logistics and schedule, (5) Quality objectives, (6) Quality control procedures, and (7) Documentation procedures. Each section comprises the following:

10.1.1.1 *Purpose of the Project*—Discusses why the test is being conducted and whether total or speciated mercury is to be determined.

10.1.1.2 *Project Organization*—Provides the personnel structure and identify team members and qualifications of the team leader with respect to Practice D7036. Identifies the analytical laboratory that will analyze the samples. Identifies the Quality Assurance Directors of both the test team and laboratory and their respective responsibilities in the test program.

10.1.1.3 *Test Logistics*—Details the test schedule, sampling locations, equipment to be used, role of sampling and plant personnel during sampling. Addresses communication mecha-

nisms within the team and with the plant. A site visit prior to the testing is recommended.

10.1.1.4 *Quality Objective*—States the quality objectives for equipment calibration, test parameters, and precision and bias of the results.

10.1.1.5 *Quality Control Procedures*—Describes quality control procedures used for (1) equipment calibration, (2) glassware cleaning and handling, (3) Chain-of-custody describing sample management from the point of collection to analysis and final data reduction (see Guide D4840), (4) isokinetic sampling, (5) field spikes, (6) sample blanks, and (7) laboratory analysis (including spikes, blanks, replicates, and calibration procedures).

10.1.1.6 *Documentation Procedures*—Includes the format of the test report, data sheet custody and integrity, and backup for electronic files.

10.1.2 *Preliminary Stack Measurements*—Select the sampling site, and determine the number of sampling points, stack pressure, temperature, moisture, dry molecular weight, and range of velocity head in accordance with procedures of Test Method D3154 or EPA Methods 1 through 4.

NOTE 7—Prior to testing, remove mercury containing devices from both staging areas and testing areas (that is, mercury manometers, broken fluorescent lamps, mercury thermometers).

10.1.3 Select the correct nozzle diameter to maintain isokinetic sampling rates based on the range of velocity heads determined in 10.1.2, and to provide adequate sample volume, without depleting the KMnO_4 .

NOTE 8—Too high of a flow rate will cause the KMnO_4 to be depleted as it reacts with SO_2 in the final set of impingers; as the KMnO_4 is depleted, it will turn brown then clear and will lose its ability to retain mercury.

10.1.4 Ensure that the proper differential pressure gage is selected for the range of velocity heads (refer to EPA Method 2, Section 2.2).

10.1.5 If the flue gas is stratified with respect to particulate concentrations, gas concentrations, or both, the stack cross-section is traversed, as specified by EPA Method 1. If the flue gas is not stratified, ample at a fixed, representative location where the flue gas is well-mixed. It is recommended that an EPA Method 17 configuration be used; however, if an EPA Method 5 setup is to be used, then select a suitable probe length such that when the stack cross-section is traversed, all traverse points can be sampled. Consider sampling from opposite sides of the stack to minimize probe length when a large duct or stack is sampled.

NOTE 9—Traversing may not be necessary, depending on the objectives of the test. In coal-fired power plants, gas-phase mercury has been found to constitute on the order of 95% of the total mercury. Errors introduced by not traversing may be either positive or negative depending upon the pattern and degree of stratification. Considering low-levels of particulate adsorbed mercury, bias introduced by sampling at a single point may be negligible. For low dust applications where the flue gas is well-mixed and a sampling location representative of the stack flow can be found, single point sampling may be adequate for obtaining representative samples.

NOTE 10—Traversing the stack may affect the performance and precision of the method if not done carefully. When traversing, the apparatus should be moved in a manner to avoid leaks or breakage of the glassware.

10.1.6 *Sampling Time and Volume*—The total sampling time for this test method should be at least 2 but not more than 3 h. Use a nozzle size that will guarantee an isokinetic gas sample volume between 1.0 dry cubic metres corrected to standard conditions (Nm^3) and 2.5 Nm^3 . If traverse sampling is done (when required), use the same points for sampling that were used for the velocity traverse as stated in 10.1.2 of this test method. Each traverse point must be sampled for a minimum of 5 min.

11. Preparation of Apparatus

11.1 Pretest Preparation:

11.1.1 Weigh several 200- to 300-g portions of silica gel in airtight containers to the nearest 0.5 g. Record the total weight of the silica gel plus container on each container. Alternatively, the silica gel can be weighed directly in the impinger immediately prior to the train being assembled.

11.1.2 Desiccate the sample filters at $20^\circ \pm 5.6^\circ\text{C}$ ($68^\circ \pm 10^\circ\text{F}$) and ambient pressure for 24 to 36 h, weigh at intervals of at least 6 h to a constant weight (that is, $<0.5\text{-}\mu\text{g}$ change from previous weighing), and record results to the nearest 0.1 μg . Alternatively, the filters may be oven-dried at 105°C (220°F) for 2 to 3 h, desiccated for 2 h, and weighed.

11.1.3 Clean all sampling train glassware as described in 8.10 before each series of tests at a single source. Until the sampling train is assembled for sampling, cover all glassware openings where contamination can occur.

11.2 Preparation of Sampling Train:

11.2.1 Assemble the sampling train as shown in Fig. 1.

11.2.2 Place 100 mL of the KCl solution (see 8.5.1 of this test method) in each of the first, second, and third impingers, as indicated in Fig. 1.

NOTE 11—For Applications with High Chloride Concentrations: Place 100 ml of the sodium thiosulfate-spiked KCl solution (see section 8.5.1) in each of the first, second and third impingers, as indicated in Fig. 1.

11.2.3 Place 100 mL of the $\text{HNO}_3\text{-H}_2\text{O}_2$ solution (see 8.5.2 of this test method) in the fourth impinger, as indicated in Fig. 1.

11.2.4 Place 100 mL of the $\text{H}_2\text{SO}_4\text{-KMnO}_4$ absorbing solution (see 8.5.3 of this test method) in each of the fifth, sixth, and seventh impingers, as indicated in Fig. 1.

11.2.5 Transfer approximately 200 to 300 g of silica gel from its container to the last impinger, as indicated in Fig. 1.

11.2.6 Prior to final train assembly, weigh and record the weight of each impinger. This information is required to calculate the moisture content of the sampled flue gas.

11.2.7 To ensure leak-free sampling train connections and to prevent possible sample contamination problems, use PTFE tape, PTFE-coated O-rings, or other non-contaminating material.

11.2.8 Place a weighed filter in the filter holder using tweezers or clean disposable surgical gloves.

11.2.9 Install the selected nozzle using a non-contaminating rubber-type O-ring or equivalent when stack temperatures are less than 260°C (500°F) and an alternative gasket material when temperatures are higher. Other connecting systems, such as PTFE ferrules or ground glass joints, may also be used on the probe and nozzle.

11.2.10 Mark the probe with heat-resistant tape or by some other method to denote the proper distance into the stack or duct for each sampling point.

11.2.11 Place crushed or cubed ice around the impingers.

11.2.12 *Leak-Check Procedures*—Follow the leak-check procedures given in Section 4.1.4.1 (Pretest Leak Check), Section 4.1.4.2 (Leak Checks During the Sample Run), and Section 4.1.4.3 (Post-test Leak Checks) of EPA Method 5 or 17. When 50 kPa vacuum is applied, the leak rate must <0.01 cfm.

NOTE 12—If O-ring seal glassware is used, the leak rate should be essentially zero (<0.01 cfm).

NOTE 13—If the flue gas temperature at the sampling location is greater than 260°C (above the temperature where PTFE or rubber-type seals can be used), the post-test leak check is determined beginning at the front end of the probe (does not include nozzle or sample filter holder for EPA Method 17).

11.3 *Preparation of the Field Blank*—A field blank is performed by assembling a sample train, transporting it to the sampling location during the sampling period, and recovering it as a regular sample. These data are used to ensure that there is no contamination as a result of the sampling activities. See 13.4.3.2. Conduct at least one field blank for each day of testing.

11.4 *Preparation of Field Spike*—A field spike is similar to the field blank, with the addition of a predetermined amount of mercury added to each of the three impinger solution. Perform the field spike by assembling a sample train, transporting it to the sampling location during the sampling period, adding the spiked solutions and recovering it as a regular sample. These solutions are then labeled and sent to the analytical lab as if they were actual samples. See 13.4.3.3. Conduct at least one field spike for each day of testing.

12. Calibration and Standardization

12.1 *Sampling Train Calibration*—Calibrate all sampling equipment, prior to the test, according to the requirements of EPA Methods referenced below and procedures detailed in the U.S. EPA Quality Assurance Handbook for Air Pollution Measurement Systems: Volume III Stationary Sources.

12.1.1 *Probe Nozzle*—Refer to Sections 6.1.1 of EPA Method 5.

12.1.2 *Pitot Tube*—Refer to Section 10 of EPA Method 2 or calibrate as in accordance with Practice D3796. Do not use Cp default value of EPA Method 2.

12.1.3 *Metering System*—Refer to Section 16.1.11 of either EPA Method for Dry Gas Meter calibration procedures. Prior to testing, perform a single-point calibration using a wet test meter or a critical orifice. The Dry Gas Meter Calibration Factor determined, must be with $\pm 2\%$ of the original value of Y to be valid.

12.1.4 *Probe Heater*—Refer to 7.1.7.1 and 7.1.7.2 of this test method and Section 10.4 of EPA Method 5.

12.1.5 *Temperature Gages*—Refer to Section 6.3 of EPA Method 2.

12.1.6 *Leak Check of the Metering System*—Refer to Section 8.4 of EPA Method 5 or Section 8.1.4 of EPA Method 17.

12.1.7 *Barometer*—Refer to Section 10.6 of EPA Method 5.

13. Procedures

13.1 *Sampling Train Operation:*

NOTE 14—A checklist for the procedures given in this section may be found in Appendix A of EPRI Report No. 1014081.

13.1.1 Maintain an isokinetic sampling rate within 10 % of true isokinetic. For an EPA Method 5 configuration, maintain sample filter exit gas stream temperatures and probe within $\pm 15^{\circ}\text{C}$ of the flue gas temperature at the sampling location. Alternatively, for reasons discussed in paragraph 16.2.3, the filter and probe may be operated at 120°C . However, regardless of the sample configuration, the sample filter, probe, or connecting umbilical cord temperature must not at any time be lower than 120°C .

13.1.2 Record the data, as indicated in Fig. 3, at least once at each sample point but not less than once every 5 min.

13.1.3 Record the dry gas meter reading at the beginning of a sampling run, the beginning and end of each sampling time increment, before and after each leak check, and when sampling is halted.

13.1.4 Level and zero the manometer. Periodically check the manometer level and zero, because it may drift during the test period.

13.1.5 Clean the port holes prior to the sampling run.

13.1.6 Remove the nozzle cap. Verify that the filter and probe heating systems are up to temperature and that the pitot tube and probe are properly positioned.

NOTE 15—For an EPA Method 5 configuration, prior to starting the gas flow through the system, the sample filter exit gas temperature may not be at the hot box temperature. However, if the system is set up correctly, once flow is established, the sample filter exit gas temperature will quickly come to equilibrium.

13.1.7 Start the pump. Position the nozzle at the first traverse point with the nozzle tip pointing in the direction of flow. Seal the openings around the probe and port hole to prevent unrepresentative dilution of the gas stream. Read the pitot tube manometer, start the stopwatch, open and adjust the control valve until the isokinetic sampling rate is obtained (refer to Section 4.1.5 from either EPA Method 5 or 17 for information on isokinetic sampling rate computations), and maintain the isokinetic rate at all points throughout the sampling period.

13.1.8 When sampling at one traverse point has been completed, move the probe to the next traverse point as quickly as possible. Close the coarse adjust valve, and shut the pump off when transferring the probe from one sample port to another. Exclude the time required to transfer the probe from one port to another from the total sampling time.

13.1.9 If the flue gas is stratified with respect to particulate concentrations, gas concentrations, or both traverse the stack cross section, as specified by EPA Method 1. If the flue gas is not stratified, sample at a fixed, representative location where the flue gas is well-mixed.

13.1.10 During sampling, periodically check and, if necessary, adjust the probe and filter exit sample gas temperatures, as well as the zero of the manometer.

13.1.11 Add more ice, if necessary, to maintain a temperature of $<20^{\circ}\text{C}$ (68°F) at the condenser/silica gel outlet.

13.1.12 Replace the filter assembly if the pressure drop across the filter becomes such that maintaining isokinetic sampling is no longer possible. Conduct a leak check (refer to EPA Method 5 or 17, Section 4.1.4.2) before installing a new filter assembly. The total particulate weight and determination of particle-bound mercury includes all filter assembly catches.

13.1.13 Monitor the color of the KMnO_4 impingers. In the unlikely event depletion of KMnO_4 by means of reduction reactions with flue gas constituents other than elemental mercury occurs, it may render it impossible to sample for the desired minimum time. This problem is indicated by the complete bleaching of the purple color of the acidified permanganate solution. If the purple color is lost in the first two H_2SO_4 - KMnO_4 impingers, then the sample must be repeated. If the gas stream is known to contain large amounts of reducing constituents (that is, >2500 ppm SO_2) or breakthrough has occurred in previous sampling runs, then the following modification is suggested: the amount of HNO_3 - H_2O_2 (10 % V/V) in the fourth impinger should be doubled, or a second HNO_3 - H_2O_2 impinger, or both should be used to increase the oxidation capacity for reducing gas components prior to the H_2SO_4 - KMnO_4 impingers.

13.1.14 Use a single train for the entire sample run, except when simultaneous sampling is required in two or more separate ducts or at two or more different locations within the same duct or when equipment failure necessitates a change of trains.

13.1.15 At the end of a sample run, turn off the coarse adjust valve, remove the probe and nozzle from the stack, record the final dry gas meter reading, and conduct a post-test leak check, as described in Section 4.1.4.3 of EPA Method 5. Also, leak-check the Pitot lines as described in EPA Method 2, Section 3.1. The lines must pass the leak check to validate the velocity head data.

13.1.16 Calculate percent isokinetic to determine whether the run was valid or another test run should be performed (refer to EPA Method 5 or 17).

13.2 *Sample Recovery*—Prepare chain-of-custody forms and sample labels prior to the test. On data sheets, provide a detailed record of custody during sampling (including all labeling information), with the initials of individuals who recover impinger contents, filters, and rinses.

NOTE 16—A checklist for the procedures given in this section may be found in Appendix A of EPRI Report No. 1014081.

13.2.1 Allow the probe to cool before proceeding with sample recovery. When the probe can be safely handled, wipe off all external particulate matter near the tip of the probe nozzle, and place a rinsed, non-contaminating cap over the probe nozzle to prevent losing or gaining particulate matter. Do not cap the probe tip tightly while the sampling train is cooling; a vacuum can form in the filter holder, with the undesired result of drawing liquid from the impingers onto the filter.

13.2.2 Before moving the sampling train to the cleanup site, remove the probe from the sampling train, and cap the open outlet. Be careful not to lose any condensate that may be present. Cap the filter inlet where the probe was fastened. Remove the umbilical cord from the last impinger, and cap the



Plant Location Operator Date Run No. Sample Box No. Meter Box No. Meter ΔH @ (kPa) C factor Pitot tube coefficient, C_p	Ambient Temperature °C (°F) Barometric Pressure kPa (in. Hg) Assumed Moisture, % Probe Length, m (ft) Nozzle Identification No. Average Nozzle Diameter, cm (in.) Probe Heater Setting °C (°F) Leak Rate, m ³ /min (cfm) Static Pressure, kPa (in. Hg) Filter No.
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Schematic of Stack Cross Section

Traverse Point Number	Sampling Time min	Vacuum kPa (in. Hg)	Stack Temperature (T _s) °C (°F)	Velocity Head (AP _s) kPa (in. H ₂ O)	Pressure Differential kPa (in. H ₂ O)	Gas Sample Volume m ³ (ft ³)	Gas Sample Temperature at Dry Gas Meter		Filter Exit Temperature °C (°F)	Probe Exit Temperature °C (°F)	Final Impinger Exit Temperature °C (°F)
							Inlet °C (°F)	Outlet °C (°F)			
Total											
Average											

FIG. 3 Mercury-Sampling Field Data Report

impinger. Cap the filter holder outlet and impinger inlet. Use non-contaminating caps, such as ground-glass stoppers, plastic caps, serum caps, or PTFE tape, to close these openings.

13.2.3 Alternatively, the following procedure may be used to disassemble the train before the probe and filter holder/oven are completely cooled. Initially disconnect the filter holder

1. Rinse filter holder and connector with 0.1N HNO₃.
2. Add 5% w/v KMnO₄ to each impinger bottle until purple color remains.
3. Rinse with 10% v/v HNO₃.
4. Rinse with a very small amount of 10% w/v NH₂OH·H₂SO₄ if brown residue remains.
5. Final rinse with 10% v/v HNO₃.

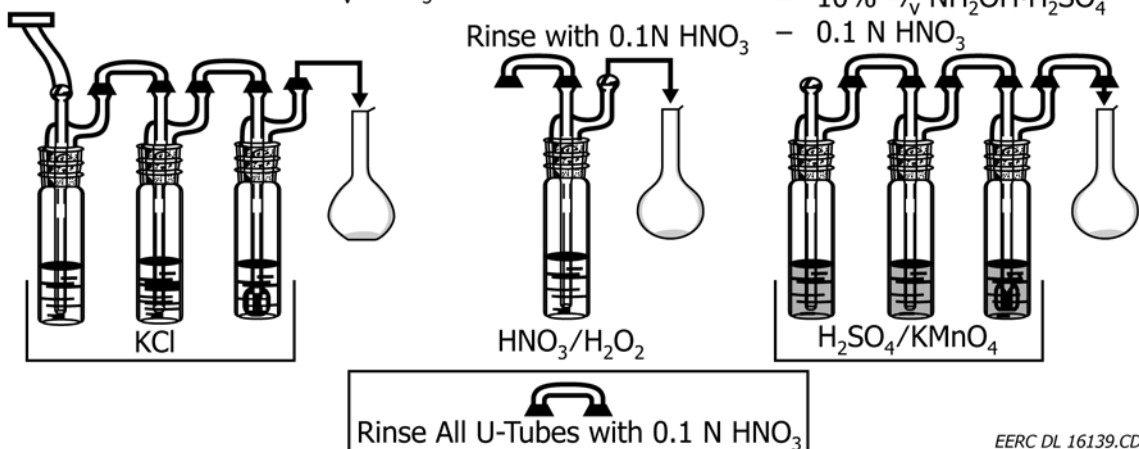


FIG. 4 Sample Recovery Scheme for the Mercury-Impinger Train

outlet/impinger inlet, and loosely cap the open ends. Then disconnect the probe from the filter holder or cyclone inlet, and loosely cap the open ends. Cap the probe tip, and remove the umbilical cord as previously described.

13.2.4 Transfer the probe and filter-impinger assembly to a clean area that is protected from the wind and other potential causes of contamination or loss of sample. Inspect the train before and during disassembly, and note any abnormal conditions.

13.2.5 The impinger train sample recovery scheme is illustrated in Fig. 4. All recovery operations should be completed within 4 h of the end of sampling activities.

NOTE 17—Impinger liquid volumes need to be kept to a minimum to improve detection limits. Probe rinses should be no more than 250 mL (Container 2), KCl samples should be no more than 500 mL (Container 3), H₂O₂/HNO₃ samples should be no more than 250 mL (Container 4), KMnO₄ samples should be no more than 500 mL (Container 5).

NOTE 18—The Table 1 summarizes the container numbers and contents used in sample recovery:

13.2.6 *Container 1 (Sample Filter)*—Carefully remove the sample filter from the filter holder so as not to lose any ash, weigh filter and ash, and place the filter in a labeled petri dish container. To handle the filter, use either acid-washed polypropylene or PTFE-coated tweezers or clean, disposable surgical gloves rinsed with water and dried. If it is necessary to fold the filter, make certain the particulate cake is inside the fold. Transfer any particulate matter or filter fibers that adhere to the filter holder gasket to the filter in the petri dish. A dry (acid-cleaned) nonmetallic bristle brush should be used to remove any remaining particulate matter. Do not use any metal-containing materials when recovering this train. Immediately cover and seal the labeled petri dish.

13.2.7 *Container 2/2a (All Rinses in Front of the Sample Filter)*:

TABLE 1 Sample Containers

Container	Description
Samples	
Container 1	Filter + Particulate matter
Container 2	Probe Rinse
Container 3	KCl impingers (combined)
Container 4	HNO ₃ -H ₂ O ₂ Impinger
Container 5	H ₂ SO ₄ -KMnO ₄ Impingers (combined)
	Desiccant (not analyzed)
Container 6	Silica gel
Blanks	
Container 7	0.1 N HNO ₃ reagent blank solution
Container 8	1N KCl reagent blank solution
Container 9	5 %v/v HNO ₃ – 10 %v/v H ₂ O ₂ reagent blank solution
Container 10	10 %v/v H ₂ SO ₄ – 4 % KMnO ₄ v/v reagent blank solution
Container 11	10 %w/v (NH ₂ OH) ₂ •HCl reagent blank solution
Container 12	Filter blank

13.2.7.1 *Case 1: Includes Gravimetric Particulate Determination in Addition to Mercury*—Quantitatively recover particulate matter and any condensate from all components prior to the sample filter. A nonmetallic brush may be used for removing particulate matter. All front-half components (all components prior to the sample filter) are then rinsed with acetone as outlined in EPA Method 5 or 17. The acetone rinse is then placed into a container (Container 2a) for which the tare weight has been recorded. Container 2a, with a ribbed watch glass over the top, is placed in a fume hood until the acetone has completely evaporated. After the front-half components have been rinsed with acetone, then rinse these components with 0.1 N HNO₃. The 0.1 N HNO₃ rinse is placed in Container 2.

13.2.7.2 *Case 2: Mercury Determination Only (No Acetone Rinse)*—Quantitatively recover particulate matter and any condensate from all components prior to the sample filter. A nonmetallic brush may be used for removing particulate matter.

The front-half components are then rinsed with 0.1 N HNO₃, and this rinse is placed in Container 2.

13.2.8 *Container 3 (Impingers 1 through 3, KCl Impinger Contents and Rinses):*

13.2.8.1 Dry the exterior surfaces of Impingers 1, 2, and 3. Then weigh and record the weight of each impinger (to the nearest 0.5 g).

13.2.8.2 Clean the filter support, the back half of the filter housing, and connecting glassware by thoroughly rinsing with 0.1 N HNO₃. Pour the rinse into a glass sample Container 3.

13.2.8.3 Add small amounts (1 mL increments) of 5 % w/v KMnO₄ solution very slowly to each KCl impinger and gently mix the impinger solution. Continue adding KMnO₄ solution until a purple color is obtained. Let the impingers sit for approximately 15 min and verify that the purple color persists. The purpose of adding KMnO₄ is to neutralize any SO₂ that may be dissolved in the KCl solution. Therefore the amount of KMnO₄ to be added depends on the SO₂ concentration in the flue gas being sampled.

NOTE 19—For application with high chloride concentrations: Add a total of 100 ml 5 % w/v KMnO₄ solution to the recovery jar and then proceed with recovery of the sodium thiosulfate-containing KCl solution as indicated in 13.2.8.3 (normal KCl solution). The purpose of adding the KMnO₄ solution is to preserve the ionic Hg. Verify that the KCl solution maintains a purple color following the addition of the KMnO₄ solution. Adding less than 100 ml of 5 % w/v KMnO₄ solution may result in clearing of the Na₂S₂O₃-spiked KCl solution during overnight storage.

13.2.8.4 Pour all of the liquid from the three KCl impingers into Container 3.

13.2.8.5 Rinse the impingers and connecting glassware with 10 % v/v HNO₃. Although unlikely, if deposits remain on the impinger surfaces, remove them by performing another 10 % v/v HNO₃ rinse that has a very small amount (several drops) of 10 % w/v hydroxylamine solution added to the HNO₃ rinse solution. Rinse each of the KCl impingers with this solution until the brown stains are removed. Add these rinses to Container 3. If the solution in Container 3 becomes clear, add a small amount of the 5 % w/v KMnO₄ solution until a pink or slightly purple color is obtained. Check again after 90 min to ensure the purple color remains.

NOTE 20—The final nitric acid concentration in the sample should be at least 1 % or greater to prevent loss of mercury from the sample.

13.2.8.6 Perform a final rinse of the impingers and connecting glassware with 0.1 N HNO₃, and add to Container 3.

13.2.8.7 Do a final rinse of all glass components with water which is discarded.

13.2.8.8 Mark the height of the fluid level in Container 3, seal, and clearly label the contents.

13.2.9 *Container 4 (Impinger 4, HNO₃-H₂O₂ Impinger Contents and Rinses):*

13.2.9.1 Dry the exterior surfaces of Impinger 4. Then weigh and record the weight of this impinger (to the nearest 0.5 g).

13.2.9.2 Pour the HNO₃-H₂O₂ absorbing solution into sample Container 4.

13.2.9.3 Rinse the H₂O₂-HNO₃ impinger and connecting glassware a minimum of two times with 0.1 N HNO₃, and pour the rinses into Container 4. Do a final rinse with water and discard water.

13.2.10 *Container 5 (Impingers 5 through 7, H₂SO₄-KMnO₄ Impinger Contents and Rinses):*

13.2.10.1 Dry the exterior surfaces of Impingers 5, 6, and 7. Then weigh and record the weight of each impinger (to the nearest 0.5 g).

13.2.10.2 Pour all of the liquid from the three H₂SO₄-KMnO₄ impingers into a glass sample Container 5.

13.2.10.3 Rinse the H₂SO₄-KMnO₄ impingers and connecting glassware a minimum of two times with 0.1 N HNO₃, and pour the rinses into Container 5. A third rinse must then be done (this rinse will remove any brown stains from the surface of the impingers). This rinse consists of 0.1N HNO₃ and several drops of 10 % w/v hydroxylamine solution (either the NH₂OH/NaCl solution or the NH₂OH · HCl solution). This rinse must have enough 10 % w/v hydroxylamine solution such that the brown stains are easily removed. If they are not easily removed, add several more drops of 10 % w/v hydroxylamine solution until the stains are completely gone. Add this rinse to Container 5. If the solution in Container 5 becomes clear, add small amounts (1 mL increments) of H₂SO₄-KMnO₄ solution until a pink or slightly purple color is obtained. Finally, preserve the solution by adding 1 mL of 5 % w/v of dichromate solution to Container 5.

13.2.10.4 Perform a final 0.1 N HNO₃ rinse of the impingers and connecting glassware followed by a water rinse. The 0.1 N HNO₃ rinse is added to Container 5, and the water rinse is discarded.

13.2.10.5 Mark the height of the fluid level, seal the container, and clearly label the contents.

NOTE 21—As stated earlier in the warning in 9.1.1, pressure can build up in the sample storage flask because of the potential reaction of KMnO₄ with acid. Do not fill the container completely, and take precautions to relieve excess pressure.

13.2.11 *Container 6 (Impinger 8, Silica Gel Impinger Contents):*

13.2.11.1 Dry the exterior surfaces of Impinger 8. Then weigh and record the weight of this impinger (to the nearest 0.5 g).

13.2.11.2 Note the color of the indicating silica gel to determine whether it has been completely spent, and make a notation of its condition. If spent, the silica gel must be either regenerated or disposed of.

13.2.12 *Solution Blanks (Containers 7-11)*—Solution blanks are taken each time new reagents are prepared.

NOTE 22—The amount of solution collected for the blanks stated below is a suggested volume.

13.2.12.1 *Container 7 (0.1 N HNO₃ Blank)*—Place 50 mL of the 0.1 N HNO₃ solution used in the sample recovery process into a properly labeled container. Seal the container.

13.2.12.2 *Container 8 (1 N KCl Blank)*—Place 50 mL of the 1 N KCl solution used as the impinger solution into a properly labeled container. Seal the container.

13.2.12.3 *Container 9 (5 % v/v HNO₃-10 % v/v H₂O₂ Blank)*—Place 50 mL of the HNO₃-H₂O₂ solution used as the nitric acid impinger reagent into a properly labeled container. Seal the container.

13.2.12.4 *Container 10 (H₂SO₄-KMnO₄ Blank)*—Place 50 mL of the H₂SO₄-KMnO₄ solution used as the

impinger solution in the sample recovery process into a properly labeled container. Refer to **Note 21** in **13.2.10.5** of this test method.

13.2.12.5 Container 11 (10 %^{w/v} Hydroxylamine Solution)—Place 100 mL of hydroxylamine solution into a properly labeled sample container. Seal the container.

13.2.13 Container 12 (Sample Filter Blank)—Once during each field test, place into a properly labeled petri dish three unused blank filters from the same lot as the sampling filters. Seal the petri dish.

13.2.14 Store the samples out of direct sunlight and complete chain-of-custody forms as applicable. It is recommended that samples be analyzed within 45 days. Report the time between sample acquisition and analysis.

13.2.15 After all impingers and connectors have been properly rinsed and the solutions recovered, the glassware should be cleaned in accordance with the procedures in **8.10** or triple-rinsed with 10 %^{v/v} HNO₃ followed by a rinsing with water. If a new source is to be sampled or if there are any brown stains on the glassware, then the glassware must be cleaned in accordance with procedures in **8.10** of this test method. If multiple sites are to be sampled during a single mobilization, an exception to this procedure will be allowed. In this case, a triple rinsing of the glassware with 10 %^{v/v} HNO₃ solution followed by a water rinse prior to sampling can be used as an alternative to the procedures in **8.10**. However, if there are any brown stains on the glassware, then the glassware must be cleaned in accordance with procedures in **8.10** of this test method.

13.3 Analytical Laboratory Procedures—Sample Preparation: Analytical laboratories conducting these procedures must be experienced with EPA Method 7470 and application of the method to this standard.

NOTE 23—Method 7470 was developed for the analyses of water samples and requires modification when applied to gas samples. The additional preparatory steps necessary for flue gas samples are given in the procedures below.

NOTE 24—Analytical laboratories may be high-volume commercial laboratories or laboratories that specialize in procedures associated with flue gas sampling method such as this test method, EPA Method 29, or both. The use of specialized laboratories experienced in this test method is recommended.

NOTE 25—It is recommended that a prepared spiked sample of each matrix type be submitted along with the collected field samples, as a quality control check.

NOTE 26—A checklist for the procedures given in this section may be found in Appendix A of EPRI Report No. 1014081.

13.3.1 Ash Sample (Containers 1 and 2):

13.3.1.1 Case 1: Includes Gravimetric Particulate Determination in Addition to Mercury—The gravimetric particulate loading is determined from the mass of the ash on the filter (Container 1) and the residual particulate from the acetone rinse (Container 2a), as outlined in EPA Method 5 or 17. If a large amount of ash is on the filter, carefully remove the ash to create a raw ash sample from which a representative-weighted aliquot can be taken for digestion. If the mass of ash collected on the filter is small (less than 0.5 g), digest the entire filter along with the ash. Dissolve the residual particulate from Container 2a using concentrated HNO₃. This solution is then added to Container 2 (0.1 N HNO₃ probe rinse). The ash

material from Container 1 is then digested using the procedures described in **13.3.2** of this test method. The same procedure is used to determine the mercury on the sample filter blank. Use a modification of EPA SW 846 7470A to digest the sample in Container 2 prior to analysis. The main modification is that the volumes of reagents and sample have been reduced tenfold to reduce waste. This reduction in reagent volume is acceptable because modern dedicated mercury analyzers do not require the large volumes that previous manual methods required. Transfer a 10-mL aliquot of the sample to a digestion tube with a screw cap.

13.3.1.2 Case 2: Mercury Determination Only—The same procedures are followed as described previously in **13.3.1.1** with the exception that there is no Container 2a.

13.3.2 Ash Digestion—Accomplish the complete dissolution of ash by one of the following methods or an equivalent alternative method. The following methods are for the dissolution of inorganic samples, such as ash or sediments, when an analysis of trace elements including mercury is done.

13.3.2.1 Microwave Digestion—The use of this test method assumes proper training in microwave digestion techniques. In addition, this test method is tailored for a specific microwave digestion system and slight modifications may therefore be required. A 0.5-g ash sample, accurately weighed to 0.0001 g, is placed in a PTFE microwave digestion vessel with 3 mL of concentrated HF, 3 mL of concentrated HNO₃, and 3 mL of concentrated HCl. The vessel is sealed and placed in the microwave (along with other vessels). The vessels are slowly heated to a pressure of 347 kPa (50 psi), which is held for 5 min, followed by heating to a pressure of 550 kPa (80 psi), which is held for 20 min. The vessels are allowed to cool to room temperature before venting. 15 mL of 4 %^{w/v} boric acid is added to each vessel. The vessels are sealed and placed in the microwave again. The vessels are slowly heated back to a pressure of 347 kPa (50 psi) and held for 10 min. The vessels are again allowed to cool to room temperature before venting. The contents of each vessel are quantitatively transferred to a 50-mL PMP or polypropylene (PP) volumetric flask and diluted; note that care must be taken in adding water to a strong acid solution.

13.3.2.2 Conventional Digestion—The use of this test method assumes proper training in PTFE bomb digestion techniques. Place a 0.5-g ash sample, accurately weighed to 0.0001 g, in a PTFE digestion vessel with 7 mL of concentrated HF and 5 mL of aqua regia. Seal the vessel, and place it in an oven or water bath at 90°C for a minimum of 8 h (these may be heated overnight). Cool the vessel to room temperature before venting. Add 3.5 g of boric acid and 40 mL of water to each vessel. Seal the vessels, and place them in the oven or water bath for an additional 1 h. Cool the vessels again to room temperature before venting. Quantitatively transfer the contents of each vessel to a 100-mL PMP, PP, or glass volumetric flask and dilute. Note that care must be taken in adding water to a strong acid solution.

13.3.3 Preparation of Aqueous KCl Impinger Solution (Containers 3 and 8)—Confirm that the sample has retained its purple color from sample recovery procedures. Absence of a purple color may be indicative of sample degradation. This fact

should be noted and reported. Clear the sample by adding 10 mL of hydroxylamine in 5 mL increments while stirring sample with a stir bar, waiting 2 min between additions. After the sample has cleared, rinse the sides of the container and lid using a transfer pipet. Transfer the cleared sample to a 500 mL volumetric flask, dilute the sample to volume with de-ionized water, and mix. If the recovered volume is greater than 500 mL then dilute the sample to 600 mL. Return the diluted sample to the jar and mix thoroughly, prior to aliquoting into digestion tubes. Use a modification of EPA SW 846 7470A to digest the sample prior to analysis. The main modification is that the volumes of reagents and sample have been reduced tenfold to reduce waste. This reduction in reagent volume is acceptable because modern dedicated mercury analyzers do not require the large volumes that previous manual methods required. Transfer a 10-mL aliquot of the sample to a digestion tube with a screw cap. Add 0.5 mL of concentrated H_2SO_4 , 0.25 mL of concentrated HNO_3 , and 10 mL of 5 %w/v KMnO_4 solution. Mix the solution, and allow it to stand for 15 min. Add 0.75 mL of 5 %w/v $\text{K}_2\text{S}_2\text{O}_8$ solution, and tightly cap the tube. Weigh the tube and record the pre-digest weight. Place the tube in a dry block heater or water bath equipped with a temperature probe, and heat to 95°C. Do not allow the temperature to exceed 95°C. Hold the sample at 95°C for 2 h before allowing it to cool to room temperature for 2 h or overnight. The purple color from the added KMnO_4 solution must remain throughout the digestion. Clearing of the solution during the heating indicates the depletion of KMnO_4 . If the solution goes clear add more 5 %w/v KMnO_4 (in 1 mL increments) to the sample until a purple color persists. Weigh the tube and record the post-digest weight. If the difference between pre-digest and post-digest weights is greater than 1 % of the pre-digest weight it should be corrected, either mathematically in the dilution factor calculation, or by adding the weight back with de-ionized water. Add 10 mL of 10 %w/v hydroxylamine solution to the sample, in 2 mL increments, waiting 30 s between each addition. Mix the solution with a transfer pipet until clear, making sure to rinse the sides of the tube and the lid. Perform the analysis immediately after clearing the sample to avoid loss of mercury. Record the volumes of the solution additions used in the preparation procedure and adjust the *DF* factor in Eq 9 accordingly.

13.3.4 Preparation of HNO_3 - H_2O_2 Impinger Solution (Containers 4 and 9)—This solution typically contains small amounts of mercury, so it should not be diluted. Instead, measure the volume with a volumetric cylinder and record. Treat the sample with a modified version of EPA SW 846 7470A. Modifications to the method are necessary to properly treat the H_2O_2 -containing impinger solution before the analysis with CVAAS or CVAFS. The modifications include the addition of HCl, the use of an ice bath during the KMnO_4 addition, and the slow addition of the KMnO_4 . Transfer a 5-mL aliquot of the sample to a digestion tube with a screw cap. Add 0.5 mL of concentrated HCl, 0.25 mL of concentrated H_2SO_4 , place the tube in an ice bath, and allow it to cool for 15 min. The destruction of H_2O_2 is accomplished by slow addition of saturated KMnO_4 solution in 0.25- or 0.50-mL increments along the inside of the digestion tube. The violence of this

reaction requires careful, slow addition of the KMnO_4 for safety reasons and to avoid loss of analyte. Cool the sample for 1 min in between additions, and mix the sample with a transfer pipet prior to each addition. Carry out the addition of KMnO_4 until the solution remains purple, indicating complete reaction of the H_2O_2 . Record the volume of saturated KMnO_4 solution added to the sample. Add 0.75 mL of 5 %w/v $\text{K}_2\text{S}_2\text{O}_8$ solution to the sample, and then cap the tube tightly. Weigh the tube and record the pre-digest weight. Place the tubes in a dry block heater or water bath equipped with a temperature probe, and heat to 95°C. Do not allow the temperature to exceed 95°C. Maintain the sample at 95°C for 2 h before allowing it to cool to room temperature for 2 h or overnight. The purple color due to KMnO_4 must remain throughout the digestion. Clearing of the solution during the heating indicates the depletion of KMnO_4 . If the solution clears, add more 5 %w/v KMnO_4 (in 1 mL increments) to the sample until a purple color persists. Weigh the tube and record the post-digest weight. If the difference between pre-digest and post-digest weights is greater than 1 % of the pre-digest weight it should be corrected, either mathematically in the dilution factor calculation, or by adding the weight back with de-ionized water. Add 10 mL of 10 %w/v hydroxylamine solution to the sample, in 2 mL increments, waiting 30 s between additions. Mix the solution with a transfer pipet until clear, making sure to rinse the sides of the tube and the lid. Perform the analysis immediately after clearing the sample to avoid loss of mercury. Record the volumes of the solution additions used in the preparation procedure and adjust the *DF* factor in Eq 13 accordingly.

13.3.5 Preparation of H_2SO_4 - KMnO_4 Impinger Solution (Containers 5 and 10)—Confirm that the sample has retained its purple color from sample recovery procedures. Absence of a purple color may be indicative of sample degradation. This fact should be noted and reported. Clear the sample by adding 30 mL of hydroxylamine in 5 mL increments while stirring sample with a stir bar, waiting 2 min between additions. After the sample has cleared, rinse the sides of the container and lid using a transfer pipet. Transfer the cleared sample to a 500 mL volumetric flask, dilute the sample to volume with de-ionized water, and mix. If the recovered volume is greater than 500 mL then dilute the sample to 600 mL. Return the diluted sample to the jar and mix thoroughly, prior to aliquoting into digestion tubes. Add the hydroxylamine slowly because of the violence of this reaction. Transfer a 10-mL aliquot of the sample to a digestion tube with a screw cap. Add 0.75 mL of 5 %w/v $\text{K}_2\text{S}_2\text{O}_8$ solution, 0.5 mL of concentrated HNO_3 , and 10.0 mL of 5 %w/v KMnO_4 solution, and tightly cap the tube. Weigh the tube and record the pre-digest weight. Mix the solution. Place the tube in a dry block heater or water bath equipped with a temperature probe, and heat to 95°C. Do not allow the temperature to exceed 95°C. Hold the sample at 95°C for 2 h before allowing it to cool to room temperature for 2 h or overnight. The purple color of the KMnO_4 solution must remain throughout the digestion. Clearing of the solution during the heating indicates the depletion of KMnO_4 . If the solution clears, add more 5 %w/v KMnO_4 (in 1 mL increments) to the sample until a purple color persists. Weigh the tube and

record the post-digest weight. If the difference between pre-digest and post-digest weights is greater than 1 % of the pre-digest weight it should be corrected, either mathematically in the dilution factor calculation, or by adding the weight back with de-ionized water. Add 10 mL of 10 %^{w/v} hydroxylamine sulfate solution to the sample, in 2 mL increments, waiting 30 s between additions. Mix the solution with a transfer pipet until clear, making sure to rinse the sides of the tube and the lid. Perform the analysis immediately after clearing the sample to avoid loss of mercury. Record the volumes of the solution additions used in the preparation procedure and adjust the *DF* factor in Eq 12 as necessary.

13.3.6 Simplification of the Digestion—If an acetone rinse was not used for gravimetric particulate determination or it is very clear that there is insignificant organic material present in the sampled gas stream; then the digestion procedure for the HNO₃–H₂O₂ and H₂SO₄–KMnO₄ impinger solutions may be simplified by omitting the persulfate digest (the addition of K₂S₂O₈ and heating). The persulfate digest is performed for the purpose of oxidizing certain organics. Because this test method is specific to coal combustion systems where organic compounds are usually insignificant⁶, this digest may be omitted because the H₂O₂ is sufficient to oxidize most compounds. The decision to omit this procedure should be made based on the gas stream being sampled or verification that organics resistant to H₂O₂ oxidation are not present, or both. If unsure whether organics are present or if an acetone rinse has been used, then the total digestion procedure is required.

13.3.6.1 Simplified Procedure for the Preparation of HNO₃–H₂O₂ Impinger Solution—If the simplified procedure can be used for the HNO₃–H₂O₂ impinger solution, the concentrated H₂SO₄ and 5 %^{w/v} K₂S₂O₈ are not added to the HNO₃–H₂O₂ aliquot sample. Also it is not necessary to heat the sample to 95°C followed by 2 h of cooling. However, it is still necessary that the concentrated HCl be added to the solution. Add 1 mL of 10 %^{w/v} hydroxylamine solution to the sample, and perform the analysis as soon as possible to avoid loss of mercury. The sample should then become clear. If the simplified procedure is used, V(K₂S₂O₈) and V(H₂SO₄) are zero when calculating *DF* in Eq 12, Section 15.

13.3.6.2 Simplified Procedure for the Preparation of H₂SO₄–KMnO₄ Impinger Solution—If the simplified procedure can be used for the H₂SO₄–KMnO₄ impinger solution, the concentrated HNO₃ and 5 %^{w/v} K₂S₂O₈ are not added to the H₂SO₄–KMnO₄ aliquot sample. Also it is not necessary to heat the sample to 95°C followed by 2 h of cooling. Add 1 mL of 10 %^{w/v} hydroxylamine solution to the sample, and perform the analysis as soon as possible to avoid loss of mercury. The sample should then become clear. If the simplified procedure is used, V(K₂S₂O₈) and V(HNO₃) are zero when calculating *DF* in Eq 13, Section 15.

13.3.7 Reagent Blanks (Containers 8 through 10)—These samples are not diluted prior to taking an aliquot. Once an

aliquot is taken, the preparation steps for each of the solutions (as well as the mercury concentration calculations) are the same as described above. These are: 13.3.3 for the aqueous KCl reagent blank, 13.3.4 for the HNO₃–H₂O₂ reagent blank, and 13.3.6.2 for the H₂SO₄–KMnO₄ reagent blank.

13.3.8 0.1 N HNO₃ and 10 %^{w/v} Hydroxylamine Rinse Solutions (Containers 7 and 11)—These solutions can be analyzed directly for mercury without any preparation steps.

13.3.9 Sample Time to Analysis—Analyze the samples immediately after digestion and no longer than one hour afterwards.

NOTE 27—Mercury can devolatilize from a digestate during the period following digestion and before analysis, resulting in low bias.

13.4 Analytical Laboratory Procedures – Sample Analysis—Analyze all of the prepared solutions by CVAAS or CVAFS following the guidelines specified by the instrument manufacturer.

NOTE 28—CVAFS is more sensitive than CVAAS and can be calibrated with lower-level standards than CVAAS.

13.4.1 Atomic Absorption or Atomic Fluorescence Spectrometer Calibration—Perform instrument setup and optimization in accordance with the manufacturer's specifications. Cold-vapor generation of mercury is performed via addition of 10 % stannous chloride in 10 % HCl solution to reduce oxidized mercury to its elemental state. The mercury-laden solution is then purged with a carrier gas into the atomic absorption cell. This procedure is used to calibrate the instrument using 3 %^{v/v} HCl as the blank along with the standards described in 8.9.3. Calibration is verified by analyzing the QC standard prepared in accordance with 8.9.4 of this test method.

13.4.1.1 Preparation of Working Hg Calibration Standards—Prepare all working standards by digesting along with the samples. Prepare digested standards of 0.25, 0.5, 1.0, 2.5, 5.0, 7.5, and 10.0 µg/L by aliquoting 100, 200, 400, 1000, 2000, 3000 and 4000 mL of the 100 µg/L Hg Stock Solution into separate digestion tubes. Bring the volume of each tube to 8.75 mL with de-ionized water. In addition, a 0.0 µg/L is prepared by adding 8.75 mL of de-ionized water to a separate tube. To each tube add 0.75 mL of 5 %^{w/v} K₂S₂O₈ solution, 0.5 mL of concentrated HNO₃, and 20.0 mL of 5 %^{w/v} KMnO₄ solution and tightly cap the tube. Weigh the tube and record the pre-digest weight. Mix the solution. Place the tube in a dry block heater or water bath equipped with a temperature probe, and heat to 95°C. Do not allow the temperature to exceed 95°C. Hold the sample at 95°C for 2 h before allowing it to cool to room temperature for 2 h or overnight. The purple color from the addition of the KMnO₄ solution must remain throughout the digestion. Weigh the tube and record the post-digest weight. If the difference between the pre-digest and post-digest weights is greater than 1 % of the pre-digest weight, it should be corrected by adding the weight back with de-ionized water. Add 10 mL of 10 %^{w/v} hydroxylamine solution to the sample in 2 mL increments, waiting 30 s between additions. Mix the solution with a transfer pipet until clear, making sure to rinse the sides of the tube and the lid. Perform the analysis immediately after clearing the sample to avoid loss of mercury.

13.4.1.2 Instrument Calibration—Analyze the standards by CVAA or CVAFS following the guidelines specified by the

⁶ "A Comprehensive Assessment of Toxic Emissions from Coal-Fired Power Plants: Phase I Results from the U.S. Department of Energy Study," Prepared for the U.S. Department of Energy Federal Energy Technology Center, Contract No. DE-FC21-93MC30097, Energy & Environmental Research Center, University of North Dakota, Grand Forks, ND, 1996.

instrument manufacturer. Construct a calibration curve by plotting the absorbances of the standards versus $\mu\text{g/L}$ Hg. The R^2 for the calibration curve should be 0.999 or better.

(1) If the curve does not have an R^2 value equal to or better than 0.999 then the curve should be rerun. If the curve still does not meet this criteria then prepare new standards and recalibrate the instrument.

(2) *Quality Control Standard*—Verify the calibration with a quality control standard prepared from a different source than that used for the preparation of the calibration standards. The reading is acceptable if it is within 95 and 105 % of the known value.

NOTE 29—Run the quality control standard immediately after the instrument is calibrated. The instrument must read within ± 5 % of the expected value before the instrument is used to analyze the test samples. A midpoint standard is then run after every 10 samples to check the slope of the calibration curve during the sample analysis.

(3) If the mercury concentration falls outside of the low or high end of the calibration curve, rerun the samples at an alternate dilution to remain within the curve.

13.4.1.3 *Sample Analysis*—Precondition the sample by passing it through a soda-lime trap and a magnesium perchlorate trap before it enters the absorption cell. Analyze the samples in duplicate following the same procedures used for instrument calibration. From the calibration curve, determine sample Hg concentrations. To determine total Hg mass in each sample fraction, refer to calculations in Section 15. Record all sample dilutions.

NOTE 30—If samples are shown to be less than 1.0 $\mu\text{g/L}$ Hg, it is recommended that sample concentrations are recalculated using a calibration curve generated using the 0.0, 0.25, 0.5 and 1.0 $\mu\text{g/L}$ standards. Most CVAA and CVAF analyzers have software that will allow you to do this without having to reanalyze the standards or samples. This is critical for achieving accurate, repeatable results in the range below 1.0 $\mu\text{g/L}$. The reason for this is that, at least for some CVAA instruments, the slope between 0.0 and 1.0 $\mu\text{g/L}$ is different than the slope between 1.0 and 10.0 $\mu\text{g/L}$.

13.4.2 *Analysis Quality Assurance/Quality Control:*

13.4.2.1 *Continued Calibration Performance*—To verify continued calibration performance, a continuing calibration check standard is run every 10 samples by reading one of the midpoint calibration standards prepared under §1.3.4.1.1. The measured mercury concentration of the continuing calibration check standard must be within 10 % of the expected value.

13.4.2.2 *Measurement Precision*—The QA/QC for the analytical portion of this test method is that every sample, after it has been prepared, is to be analyzed in duplicate with every tenth sample analyzed in triplicate. These results must be within 10 % of each other. If this is not the case, then the instrument must be recalibrated and the samples re-analyzed.

NOTE 31—It is important to clarify to the analytical laboratory that the procedures of this standard are to be followed with respect to analyzing replicates. Commercial laboratories may have quality control procedures where duplicates are not performed on every sample or triplicates are performed less frequently. Also, commercial laboratories may batch samples from multiple clients, which may result in triplicate analyses not being conducted for the test submitted. Triplicate samples are to be conducted on samples obtained during the relevant test.

NOTE 32—Laboratories having quality control specifications less stringent than those of this standard must be informed of the requirements of

the standard for multiple analysis and that corrective action must be taken when the criteria are not met, as in accordance with this section (13.4.2.2).

13.4.2.3 *Measurement Accuracy*—Following calibration, an independently prepared standard (not from same calibration stock solution) must be analyzed (§1.3.1.4.1.2.2). In addition, each sample matrix (standard addition) at a frequency of one per batch or 1 out of ten, (10 %), whichever is greater, and analyze. The measured mercury content of the spiked samples must be within 10 % of the expected value.

NOTE 33—Each sample that is selected for triplicate analyses is split into four samples. The first three are analyzed to determine the standard deviation between samples. The fourth sample is spiked with a known amount of mercury and the results then compared to the average of three samples to determine the spike recovery.

13.4.2.4 *Independent QA/QC Checks*—For the ash samples, a certified reference ash sample (may be purchased from NIST) is to be digested and analyzed at least once during the test program. It is also suggested that the QA/QC procedures developed for a test program include submitting, on occasion, spiked mercury samples to the analytical laboratory by either the prime contractor, if different from the laboratory, or an independent organization. The measured mercury content of reference samples must be within 15 % of the expected value. If this limit is exceeded, corrective action (for example, re-calibration) must be taken and the samples re-analyzed.

13.4.3 *QA/QC*—For this test method, it is important that both the sampling team and analytical personnel be very well trained in the procedures. This is a complicated method that requires a high-level of sampling and analytical experience. For the sampling portion of the QA/QC procedure, both solution and field blanks are required. It should be noted that if high-quality reagents are used and care is taken in their preparation and in the train assembly, there should be little, if any, mercury measured in either the solution or field blanks.

13.4.3.1 *Solution Blanks*—As stated in 13.2.12 of this test method, solution blanks will be taken and analyzed every time a new batch of solution is prepared. If mercury is detected in these solution blanks, the concentration is subtracted from the measured sample results. The maximum amount that can be subtracted is 10 % of the measured result or 10 times the detection limit of the instrument whichever is lower. If the solution blanks are greater than 10 % the data must be flagged as suspect.

13.4.3.2 *Field Blanks*—A field blank is performed by assembling a sample train, transporting it to the sampling location during the sampling period, and recovering it as a regular sample. These data are used to ensure that there is no contamination as a result of the sampling activities. A minimum of one field blank at each sampling location must be completed for each test site. Any mercury detected in the field blanks cannot be subtracted from the results. Whether or not the mercury detected in the field blanks is significant is determined based on the QA/QC procedures established prior to the testing. At a minimum, if field blanks exceed 30 % of the measured value at the corresponding location, the data must be flagged as suspect.

13.4.3.3 *Field Spikes*—A field spike is similar to the field blank, with the addition of a predetermined amount of mercury

TABLE 2 Data Quality Objectives

Measure	Sample Type	Objective	Approach/Frequency
Accuracy	Field spike	<±15 % of true value	Collect and analyze one field spike each unit per day of testing (analysis must be done on each solution)
Accuracy	Laboratory method and instrument spikes	<±10 % of true value	One per batch per solution type or 10 %, whichever is greater
Accuracy	Reference material	8 5% – 115 % of reference value	One per test program
Precision	Duplicate analyses	Standard deviation <0.5 µ/L	One per batch per solution type or 10 %, whichever is greater
Precision	Triplicate analyses	Standard deviation <0.5 µ/L	One per batch per solution type or 10 %, whichever is greater
Contamination	Reagent blank	<0.1 µ/L	One blank per batch of each reagent
Contamination	Field blank	Maximum of 0.5 µ/L g/L and <10 % of sampled mercury concentration	Collect and analyze one field blank for each unit per day of testing (analysis must be done on each solution)

added to each of the three impinger solutions. Perform the field spike by assembling a sample train, transporting it to the sampling location during the sampling period, adding the spiked solutions and recovering it as a regular sample. A minimum of one field spike at each sampling location, for each day of testing is required. Spike recoveries should be between 85 and 115 % of the known value, or the data should be flagged as suspect.

13.4.4 *Summary-Data Quality Objectives*—Table 2 summarizes the data quality objective for the preparation and analysis of flue gas mercury samples in this test method.

13.4.4.1 *Failure to Meet Quality Objectives*—If the objectives of Table 2 are not met, repeat the analysis or retest if necessary. If the test(s) cannot be repeated and a data quality objective is not met, flag any affected data and provide a discussion and an assessment of the validity of the data in the final data report.

13.4.5 *Reporting*—The quality control results for all replicates, blanks, and spikes are to be included in the final data report issued by the laboratory. Report also, the Method Detection Limit (MDL).

14. Flue Gas Calculations

14.1 *Dry Gas Volume*—Calculate the dry gas sample volume, $V_{m(std)}$, at standard conditions using Eq 1.

$$V_{m(std)} = V_m Y \left(\frac{T_{std}}{T_m} \right) \left[\frac{P_{bar} + \Delta H}{P_{std}} \right] = K_I V_m Y \frac{P_{bar} + \Delta H}{T_m} \quad (1)$$

where:

- P_{bar} = barometric pressure at the sampling site, kPa (in. Hg),
- P_{std} = standard absolute pressure, 101.3 kPa (29.92 in. Hg),
- T_m = absolute average dry gas meter temperature (refer to Fig. 3), K (°R),
- T_{std} = standard absolute temperature, 293 K (528°R),
- V_m = volume of gas sample as measured by dry gas meter, m³ (dscf),

$V_{m(std)}$ = volume of gas sample measured by the dry gas meter, corrected to standard conditions, Nm³ (dscf),

Y = dry gas meter calibration factor,

ΔH = average pressure differential across the orifice meter (refer to Fig. 3), kPa (in. Hg), and

K_I = 2.894 K/kPa (17.64°R/in. Hg).

NOTE 34—Eq 1 can be used as written unless the leakage rate observed during any of the mandatory leak checks (that is, leak checks conducted prior to component changes or following the test) exceeds the maximum acceptable leakage rate, L_a , equal to 0.00057 m³/min (0.02 cfm) or 4 % of the average sampling rate, whichever is less. If the leakage rate observed during the post-test leak check, L_p , or an individual leakage rate observed during the leak check conducted prior to the “ith” component change ($i = 1, 2, 3, \dots, n$), L_i , exceeds L_a , then Eq 1 must be modified as follows:

Case I—No component changes made during sampling run. In this case, replace V_m with the expression:

$$[V_m - (L_p - L_a)\theta]$$

where:

L_p = leakage rate observed during the post-test leak check, m³/min (cfm),

L_a = maximum acceptable leakage rate for either a pretest leak check or for a leak check following a component change—equal to 0.00057 m³/min (0.02 cfm) or 4 % of the average sampling rate, whichever is less, and

θ = total sampling time, min.

Case II—One or more component changes made during the sampling run. In this case, replace V_m with the expression:

$$\left[V_m - (L_i - L_a)\theta_1 - \sum_{i=1}^n (L_i - L_a)\theta_i - (L_p - L_a)\theta_p \right]$$

where:

θ_i = sampling time interval, from the beginning of a run until the first component change, min.

and substitute only for those leakage rates (L_i or L_p) that exceed L_a .

14.2 *Volume of Water Vapor*—Calculate the volume of water vapor of the stack gas using Eq 2.

$$V_{w(std)} = \frac{W_{lc}RT_{std}}{M_w P_{std}} = K_2 W_{lc} \quad (2)$$

where:

- M_w = molecular weight of water, 18.0 g/g-mole (18.0 lb/Lb-mole),
 R = ideal gas constant, 0.008314 kPa-m³/K-g-mole (21.85 in. Hg-ft³/°R-lb-mole),
 W_{lc} = total weight of liquid collected in impingers and silica gel (refer to Fig. 3), g,
 $V_{w(std)}$ = volume of water vapor in the gas sample, corrected to standard conditions, m³ (scf), and
 K_2 = 0.001336 m³/mL (0.04707 ft³/mL).

14.3 *Volume of Moisture*—Calculate the moisture content, B_{ws} , of the stack gas using Eq 3.

$$B_{ws} = \frac{V_{w(std)}}{V_{m(std)} + V_{w(std)}} \quad (3)$$

where:

- B_{ws} = water vapor in the gas stream, proportion by volume.

15. Calculations for Particle-Bound, Oxidized, Elemental, and Total Mercury Concentrations

15.1 Particle-Bound Mercury:

15.1.1 *Case 1: Amount of Ash on the Filter is Greater Than 0.5 g*—Calculate the concentration of mercury in µg/g in the ash sample (Hg_{ash}) using Eq 4:

$$Hg_{ash}, \mu g/g = (IR)(DF) \quad (4)$$

where:

- IR = instrument reading, µg/L, and
 DF = dilution factor = (total digested volume, L)/(mass of ash digested, g)

$$Hg_{pr}, \mu g = (IR)(V_1) \quad (5)$$

where:

- IR = instrument reading, µg/L, and
 V_1 = total volume of probe rinse sample from which sample aliquot was taken, L.

Eq 5 assumes no preparation steps are needed prior to analyzing the probe rinse for mercury using CVAA. Although not required, a persulfate digest can be done on the probe rinse sample as discussed in 13.3.3. If the persulfate digest can be done, Eq 5 becomes $Hg_{pr}, \mu g = (IR)(V_1)DF$ where DF is the same as Eq 9. There is no filter blank subtraction when >0.5 g of ash are collected on the sample filter or thimble. The amount of particle-bound mercury (Hg_{fb}) is then determined using Eq 6:

$$Hg \text{ (particle)}, \mu g = (Hg_{ash})(W_{ash}) + Hg_{pr} \quad (6)$$

where:

- W_{ash} = total mass of ash on filter, g.

The concentration of particle-bound mercury (µg/Nm³) in the gas stream is then determined using Eq 7:

$$Hg^{pp}, \mu g/Nm^3 = Hg \text{ (particle)} / V_{m(std)} \quad (7)$$

where:

- $V_{m(std)}$ = total volume of dry gas sampled at standard (normal) conditions, Nm³.

15.1.2 *Case 2: Amount of Ash on the Filter is Less Than 0.5 g*—The calculation is the same as in Case 1 except the entire sample (ash and filter) is digested; therefore, DF in Eq 4 is defined only by the total digested volume. In addition, a filter blank is subtracted as calculated in Eq 8.

$$Hg_{pp}, \mu g = (IR)(V_2) \quad (8)$$

where:

- IR = instrument reading, µg/L, and
 V_2 = total volume of sample filter blank digest, L.

Eq 7 for Case 2 then becomes: $Hg \text{ (particle)}, \mu g = (Hg_{ash} - Hg_{fb}) + Hg_{pr}$.

15.2 Oxidized Mercury:

15.2.1 *KCl Solution (Impingers 1-3)*—Calculate the concentration of mercury in µg/L in the KCl impinger solutions using Eq 9:

$$Hg_{KCl}, \mu g/L = (IR)(DF) \quad (9)$$

where:

- IR = instrument reading, µg/L,
 DF = dilution factor,

$$DF = \frac{V_D + V(H_2SO_4) + V(HNO_3) + V(KMnO_4) + V(K_2S_2O_8) + V(NH_2OH)}{V_D}$$

- V_D = total digested volume, 10 mL,
 $V(H_2SO_4)$ = volume of added concentrated H₂SO₄, 0.5 mL,
 $V(HNO_3)$ = volume of added concentrated HNO₃, 0.5 mL,
 $V(KMnO_4)$ = volume of added 5 %w/v KMnO₄, 1.5 mL,
 $V(K_2S_2O_8)$ = volume of added 5 %w/v K₂S₂O₈, 0.75 mL, and
 $V(NH_2OH)$ = volume of added 10 %w/v hydroxylamine sulfate, 1.0 mL.

The amount of mercury in the KCl solution blank is calculated in the same way.

15.2.2 *Total Oxidized Mercury (Hg_O)* is defined by method as the mercury measured in the KCl sample minus the mercury measured in the KCl solution blanks, as shown in Eq 10:

$$Hg_O, \mu g = (Hg_{KCl})(V_3) - (Hg_{Ob})(V_4) \quad (10)$$

where:

- Hg_{KCl} = mercury concentration measured in KCl aliquot, µg/L,
 V_3 = total volume of aqueous KCl from which sample aliquot was taken, L,
 Hg_{Ob} = mercury concentration measured in KCl solution blank aliquot, µg/L, and
 V_4 = volume of aqueous KCl originally charged to the impingers, L.

TABLE 3 Results from Formal EPA Method 301 Evaluation Tests for the Ontario Hydro Method^A

Ontario Hydro Method ^B	Total Vapor-Phase Mercury			Oxidized Mercury			Elemental Mercury		
	Mean, µg/ Nm ³	Std. Dev.	RSD, %	Mean, µg/ Nm ³	Std. Dev.	RSD, %	Mean, µg/ Nm ³	Std. Dev.	RSD, %
Baseline	23.35	2.05	8.79	21.24	2.13	10.02	2.11	0.65	30.69
Hg ⁰ Spike (15.0 µg/Nm ³)	38.89	2.00	5.13	23.32	2.08	8.94	15.57	1.09	6.97
HgCl ₂ Spike (19.9 µg/Nm ³)	42.88	2.67	6.23	40.22	2.87	7.14	2.66	0.89	33.31

^A For each mean result, there were 12 replicate samples (four quadrants).

^B The correction factor in all cases was not statically significant and is not shown.

The concentration of Hg²⁺ (µg/Nm³) in the gas stream is then determined using Eq 11:

$$Hg^{2+}, \mu g/Nm^3 = Hg_o/V_{m(std)} \quad (11)$$

where:

$V_{m(std)}$ = total volume of dry gas sampled at standard conditions, m³.

15.3 Elemental Mercury:

15.3.1 *HNO₃-H₂O₂ Solution (Impinger 4)*—Calculate the concentration of mercury in µg/L in the HNO₃-H₂O₂ impinger solution using Eq 12:

$$Hg_{H_2O_2}, \mu g/L = (IR)(DF) \quad (12)$$

where:

IR = instrument reading, µg/L,
 DF = dilution factor

$$= \frac{V_D + V(HCl) + V(H_2SO_4) + V(KMnO_4) + V(K_2S_2O_8) + V(NH_2OH)}{V_D}$$

V_D = total digested volume, 5 mL,
 $V(HCl)$ = volume of added concentrated HCl, 0.25 mL,
 $V(KMnO_4)$ = volume of added saturated KMnO₄, mL (volume need to turn sample to a purple color),
 $V(K_2S_2O_8)$ = volume of added 5 %^{w/v} K₂S₂O₈, 0.75 mL (if used), and
 $V(NH_2OH)$ = volume of added 10 %^{w/v} hydroxylamine sulfate, 1.0 mL.

The concentration of mercury in the HNO₃-H₂O₂ solution blank is calculated in the same way.

15.3.2 *H₂SO₄-KMnO₄ Solution (Impingers 5-7)*—Calculate the concentration of mercury in µg/L in the H₂SO₄-KMnO₄ impinger solutions using Eq 13:

$$\text{Mercury, } \mu g/L = (IR)(DF) \quad (13)$$

where:

DF = dilution factor

$$= \frac{V_D + V(HNO_3) + V(K_2S_2O_8) + V(NH_2OH)}{V_D}$$

IR = instrument reading, µg/L,
 V_D = total digested volume, 5 mL,
 $V(HNO_3)$ = volume of added concentrated HNO₃, 0.5 mL, and
 $V(K_2S_2O_8)$ = volume of added 5 %^{w/v} K₂S₂O₈, 0.75 mL.

The concentration of mercury in the H₂SO₄-KMnO₄ solution blank is calculated in the same way.

15.3.3 *Total Elemental Mercury (Hg_E)* is defined by the method as the mercury measured in the H₂SO₄-KMnO₄ impingers plus the mercury in the HNO₃-H₂O₂ impingers minus the solution blanks as shown in Eq 14:

$$Hg_E, \mu g = (Hg_{H_2O_2})(V_4) - (Hg_{Eb_1})(V_5) + (Hg_{KMnO_4})(V_6) - (Hg_{Eb_2})(V_7) \quad (14)$$

where:

$Hg_{H_2O_2}$ = mercury concentration measured in HNO₃-H₂O₂ aliquot, µg/L,
 V_4 = total volume of aqueous HNO₃-H₂O₂ from which sample aliquot was taken, L,
 V_5 = total volume of aqueous HNO₃-H₂O₂ originally charged to the impinger, L,
 Hg_{Eb_1} = mercury concentration measured in HNO₃-H₂O₂ solution blank aliquot, µg/L,
 Hg_{KMnO_4} = mercury concentration measured in H₂SO₄-KMnO₄ aliquot, µg/L,
 V_6 = total volume of aqueous H₂SO₄-KMnO₄ from which sample aliquot was taken, L,
 V_7 = total volume of aqueous H₂SO₄-KMnO₄ originally charged to the impingers, L, and
 Hg_{Eb_2} = mercury concentration measured in H₂SO₄-KMnO₄ solution blank aliquot, µg/L.

The concentration of Hg²⁺ (µg/Nm³) in the gas stream is then determined using Eq 15:

$$Hg^0, \mu g/Nm^3 = Hg_E/V_{m(std)} \quad (15)$$

where:

$V_{m(std)}$ = total volume of dry gas sampled at standard conditions, Nm³.

15.4 *Total Mercury* is defined by the method as the sum of the particulate bound mercury, oxidized mercury, and elemental mercury as shown in Eq 16:

$$Hg(\text{total}), \mu g/Nm^3 = Hg^{pp} + Hg^{2+} + Hg^0 \quad (16)$$

16. Precision and Bias

16.1 Precision:

16.1.1 Formal evaluation of the Ontario Hydro method was completed with dynamic spiking of Hg⁰ and HgCl₂ into a flue gas stream.⁷ The results are shown in Table 3. The relative standard deviation for gaseous elemental mercury and oxidized mercury was found to be less than 11 % for mercury concentrations greater than 3 µg/Nm³ and less than 34 % for mercury

⁷ EPRI, U.S. DOE NETL. "Evaluation of Flue Gas Mercury Speciation Methods," EPRI TR-108988, Electric Power Research Institute, Palo Alto, CA, Dec. 1997.

concentrations less than $3 \mu\text{g}/\text{Nm}^3$. In all cases, the laboratory bias for these tests based on a calculated correction factor was not statistically significant. These values were within the acceptable range, based on the criteria established in EPA Method 301 (% RSD less than 50 %).

16.1.2 *Caution*—The precision of particle-bound, oxidized, and elemental mercury sampling method data is influenced by many factors: flue gas concentration, source, procedural, and equipment variables. Strict adherence to the method is necessary to reduce the effect of these variables. Failure to assure a leak-free system, failure to accurately calibrate all indicated system components, failure to select a proper sampling location, failure to thoroughly clean all glassware, and failure to follow prescribed sample recovery, preparation, and analysis procedures can seriously affect the precision of the results. Nevertheless, if suitable care is taken, a precision of 10 to 15 % is achievable.

16.2 Bias:

16.2.1 *Total Mercury Measurement*—Other than sample stability and preservation concerns, there are no known biases affecting the total mercury measurement.

16.2.2 Mercury Speciation:

16.2.2.1 Particle-bound mercury existing in the flue gas may vaporize after collection in the front half of the sampling train because of continued exposure to the flue gas sample stream and reduced pressures during the sampling period. Such vaporization would result in a negative particle-bound mercury bias.

16.2.2.2 Certain types of fly ash can catalyze oxidation of elemental mercury, causing a positive bias of the oxidized mercury and a negative bias of elemental mercury. Increases in filter temperature increase catalytic effects. These biases can be identified by speciation measurements on mercury adsorbed on the ash collected by an electrostatic precipitator or baghouse.

16.2.2.3 It is also possible that some fly ashes may adsorb mercury at the filter resulting in a low bias for the gaseous mercury. Adsorption typically increases at lower filter temperatures.

16.2.2.4 However, these behaviors are only expected where particle concentrations are high, and are therefore unlikely to be relevant at stack outlets following particulate control devices such as electrostatic precipitators or baghouses. Note that none of these conditions affect the total mercury measurement.

16.2.2.5 In combustion gases where chlorine gas (Cl_2) is present, under certain conditions, the Cl_2 may react in the liquid phase to oxidize Hg^0 . Hypochlorite ion (OCl^-), formed during the dissociation of Cl_2 in aqueous solutions, oxidizes Hg^0 to soluble $2+$. This problem can be mitigated by the addition of sodium thiosulfate ($\text{Na}_2\text{S}_2\text{O}_3$), which is specified in this standard in cases where Cl_2 is present without sufficient levels of SO_2 . The presence of sulfur dioxide (SO_2) mitigates this bias in a similar manner to the addition of sodium thiosulfate. As a result, this speciation bias is not likely to be a factor for coal combustion applications.

17. Keywords

17.1 air toxics; mercury; sampling; speciation

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